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TREATMENT OF CHEESE PROCESSING
WASTEWATERS IN AERATED LAGOONS

Francis R. Daul

Kent Cheese Company

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16. Abstract A full-scale treatment of wastewater from a cheese processing operation
was conducted over a one year period. A two-stage aerated lagoon system provided
an average BOD removal of 97 percent at BOD loadings ranging from 0.117 to 4.34
lb/1000cu ft/day. Hydraulic detention time varied from 50 to 82 days per lagoon.
Temperature had the single greatest influence on process efficiency. Costs for
lagoon operation, maintenance and amortization were estimated to be \$2.15 per
1000 gallons or \$0.14 per pound BOD applied.

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TREATMENT OF CHEESE PROCESSING WASTEWATERS
IN AERATED LAGOONS

By

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Project 12060 EKQ
Program Element 1BB037
Roap/Task 21 BAD 26

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This report has been reviewed by the Environmental Protection Agency and approved for publication. Approval does not signify that the contents necessarily reflect the views and policies of the Environmental Protection Agency, nor does mention of trade names or commercial products constitute endorsement or recommendation for use.

ABSTRACT

The treatment of cheese processing wastewaters in two-stage aerated lagoons was evaluated over a one year period. Aeration was provided by Polcon Corporation subsurface aerators (Helixors). The aeration system provided an average standard transfer rate of 4.0 lbs oxygen per kilowatt-hour at 20°C and 0.0 dissolved oxygen in tap water. Oxygen dispersion throughout the primary lagoon was normally adequate at a power input of 8.5 horsepower per million gallons. During the spring, however, dissolved oxygen concentrations in the primary lagoon were zero probably due to the solubilization of benthal deposits accumulated but not stabilized during the winter months. There was not sufficient power, however, to prevent suspended solids deposition. Horizontal velocity components varied from 1.0 fps within the central portion of the lagoon to less than 0.1 fps in the sloped peripheral zones. Over the one year period of this study, sludge accumulations ranged from 0 to 2 inches depending upon location.

The lagoon system performance during the one year period was correlated with temperature. The average total BOD removal was 97 percent producing an effluent BOD concentration of 52 mg/l. During the summer months secondary lagoon effluents were less than 20 mg/l, 70 percent of the time. Approximately 50 percent of the total effluent BOD was soluble. Effluent suspended solids were high, averaging 108 mg/l. Greater than 90 percent of the suspended solids were volatile. Total and fecal coliform reductions were normally greater than 99.9 percent throughout the study period.

During the early spring primary lagoon dissolved oxygen concentrations dropped to zero and remained below 0.1 mg/l until mid July. Benthal deposits, accumulated during the winter period, were believed to account for the unusually high oxygen demands during this period. Demands in excess of three times greater than predicted by the BOD load to the lagoon were noted.

Costs for lagoon operation, maintenance, and amortization were estimated to be \$13,377 per year, \$2.15 per 1000 gallons, \$0.14 per lb of BOD applied, and \$0.0033 per lb cheese produced.

This report was submitted in fulfillment of project 12060EKQ under the partial sponsorship of the Environmental Protection Agency.

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SECTION I

CONCLUSIONS

The conclusions summarized below were the result of the analysis of data collected over a one year period from the two-stage aerated lagoons treating cheese processing wastewaters for the Kent Cheese Company.

Each of the two lagoons were of equal volume and provided detention times of from 50 to 82 days each during the first year of study. The average flow to the lagoons was approximately 17,000 gpd, ranging from 7,300 gpd to 35,500 gpd. Raw influent BOD loading averaged 285 lbs/d ranging from less than 15 lb/d to 554 lb/d resulting in a primary lagoon loading of from 0.117 to 4.34 lb BOD/1000 cu ft/d.

1. Aerator Performance

- a. The oxygen transfer efficiency of the Polcon Corporation Helixors in the primary lagoon was 4.0 lbs oxygen/kw-hr or 3.0 lbs oxygen/hp-hr, for standard conditions (20°C, tap water, zero dissolved oxygen) at an air flow rate of 14.8 scfm per unit.
- b. An increase of the air flow rate by a factor of two per unit substantially reduced the oxygen transfer efficiency.
- c. The dispersion of oxygen in the primary lagoon at a power input of 8.5 hp/MG was nearly uniform.
- d. At the power input of 8.5 hp/MG, there was a tendency for suspended solids to accumulate in the peripheral regions where horizontal velocity components were normally less than 0.1 fps. The one year study was not long enough to evaluate the rate of solids accumulation within the lagoon.
- e. The relative oxygen transfer coefficient in the primary lagoon was approximately 81 percent of that expected in tap water. The oxygen saturation value in the primary lagoon content was approximately 98 percent of that in pure water.

2. Lagoon Performance

- a. An average of 97 percent removal of BOD was achieved in the two stage aerated lagoon system with the poorest performance (95 percent) occurring during the winter season. The average final effluent BOD concentration was 52 mg/l of which 27 mg/l was soluble. Probability distributions for treatment performance were prepared for the one year study.

- b. Temperature had an important effect on lagoon performance.
- c. Suspended solids removals in the primary lagoon were sporadic. The overall average removal of suspended solids in the lagoon system was 82 percent with an average concentration of 108 mg/l in the final effluent. Over 90 percent of the suspended solids were volatile.
- d. Approximately 29% of the volatile suspended solids in the final effluent contributed to the effluent BOD. A similar relationship for the primary effluent suspended solids indicated that 46 percent of the volatile suspended solids contributed to the BOD.
- e. High oxygen uptake rates occurring in the early spring resulted in zero D.O. concentrations in the primary lagoon. These high uptake rates are likely due to the solubilization of solids deposited during the winter months. The cycling of sludge accumulation during the cold periods and active biological stabilization of the benthic deposits during the warmer months will result in dynamic fluctuations in oxygen requirements and sludge deposits. Aerator sizing to account for these fluctuations must be provided. In this study oxygen demands in excess of three times the influent BOD were measured during the early spring.
- f. The staging of the wastewater treatment lagoons provided considerable attenuation of the fluctuating BOD and solids concentrations in the primary lagoon effluent.
- g. Total and fecal coliform reductions exceeded 99.9 percent in the two stage lagoon system without disinfection.
- h. Nitrogen concentrations in the raw wastewater were low with respect to carbon. The BOD to Nitrogen ratio was 100:0.7 indicating a substantial deficiency in nitrogen. Phosphorus concentrations were high resulting in a BOD:P ratio of 100:2.2.
- i. Total phosphorus removals of 52 percent observed during the one year study may be misleading insofar as a steady state with respect to phosphorus transformations has probably not been established. Total organic nitrogen removal of 47 percent were observed. Nitrate concentrations in the final effluent were highest during the warm summer months but were less than 0.4 mg/l during the rest of the year.

3. Costs

The cost for wastewater treatment including operation, maintenance, & capital cost amortization during the first year of operation was \$2.15 per 1000 gal., \$0.14 per lb BOD applied, and \$0.0033 per lb of cheese produced.

SECTION II

RECOMMENDATIONS

Based on the operation of the two stage cheese processing wastewater lagoons at Kent, Illinois, the following recommendations have been proposed:

1. Efforts should be made to further reduce the settleable solids input to the lagoons. These solids cause clogging problems at the inverted siphon, interfere with flow measurement and sampling at the primary lagoon, and result in increased BOD loading to the primary lagoon. Insofar as solubilization of settleable solids during the warm months has resulted in severe depressions in dissolved oxygen concentrations in the primary lagoon, every measure taken to reduce influent settleable solids will help to alleviate that problem. It is suggested that a settling tank or imhoff tank be provided ahead of the primary lagoon for this purpose.
2. The high air pressures observed during dual compressor operation suggests that the air headers and inlet orifices are undersized for the increased air flow rates. Since oxygen uptake rates during the spring and early summer exceed current oxygen transfer rates, it is recommended that investigations be made to determine the necessity for enlarging existing air piping or providing additional Helixors in the primary lagoon.
3. The increased oxygen uptake rates in facultative aerated lagoons during the spring have been reported by a number of investigators. Further investigation should be conducted to provide a quantitative estimate of this increased demand. Such a study would require data collection over a number of years at the existing aerated lagoon site.
4. Nitrogen deficiencies in the raw wastewater will normally result in poorer performance of the biological system. It is recommended that additional nitrogen be added as ammonia or urea to more closely approximate the BOD to Nitrogen ratio of 100 to 5. Comparisons in lagoon performance should be noted during this period (at least one full year) so as to evaluate the value of this supplementation.
5. The secondary lagoon effluent structure consists of only an upturned 4 inch cast iron pipe elbow. It is recommended that this effluent structure be modified so as to provide some baffling to eliminate gross solids and scum entrainment.

SECTION III

INTRODUCTION

The evaluations and findings reported herein were supported by EPA Project No. 12060 EKQ for post-construction studies over a 12-month period to demonstrate and evaluate the use of aerated lagoons for the treatment of cheese wastes. Initially the objectives of the study included the evaluation and demonstration of the effectiveness of reverse osmosis for the reduction of BOD of cheese whey wastes. The decision was made that the reverse osmosis method would not be included in this study.

The authors of this report were not engaged to design or size the system employed, select equipment utilized or were consulted regarding the process flowsheet applicable to this wastewater treatment scheme. The authors were retained to conduct the post construction study, direct frequency of sampling, recommend analyses to be performed, and evaluate and present findings of the performance of the aerated lagoon systems for the wastes received at the treatment site under the limitations of the existing budget.

The principal objectives of this study were to demonstrate the performance of a staged aerated lagoon treatment plant utilizing the Helixor* type of submerged aeration system treating cheese processing wastes over a 12 month period of operation. Part of the evaluation was directed to the performance of the aeration equipment employed wherein mixing effectiveness in terms of liquid velocities produced, uniformity of dissolved oxygen levels and sludge accumulations could be ascertained under the conditions of operation and geometric configuration of the lagoons employed. In addition the oxygen transfer efficiency of the aeration system employed was evaluated under field conditions. A major part of this study was to determine the BOD removal rate functions for each stage of aerated lagoon treatment noting the influence of BOD loadings, temperature and seasonal variations in loading and performance obtained for this type of biological treatment. The theoretical and applied concepts of biological treatment related to low-solids aerated lagoon systems were employed to evaluate the treatment system. Lastly, the performance of this type of treatment was evaluated in terms of costs associated with the wastewater characteristics and the pounds of cheese produced.

*subsurface aeration unit manufactured by Polcon Corporation, Montreal, Canada

SECTION IV

EXPERIMENTAL SYSTEM

The wastewater treatment system consisted of two equal volume aerated lagoons with a staged flowsheet wherein the effluent from Lagoon No. 1 is passed to Lagoon No. 2 before final discharge of the treated effluent. The lagoons had earth embankments with side slopes of 3 horizontal to 1 vertical, a water depth of 12 feet at the lowest central portion of the lagoon, and a length-width measurement of 157 feet by 123 feet at the water line for a volume of 955,000 gallons each (Figure 1). The first lagoon was provided with thirteen 6 foot long 18 inch diameter Helixors arranged in a pattern of 2 rows of 4 units equally spaced along the intersection of the flat bottom and side slopes of the long dimension of the lagoon. Five additional units were added to this lagoon between the 2 rows of four units such that 4 of the five units were equally spaced in 2 rows between the 1st and 2nd and 2nd and 3rd aeration units. The fifth additional aerator was equally spaced between the third and fourth unit of the initial eight unit arrangement (Figure 1). The second lagoon was provided with three aeration units arranged in a triangular pattern close to the inlet end of the lagoon in the lowest central section of the lagoon.

A separate blower building with two rated 240 cfm Gardner Denver rotary blowers (one standby), Model 3CDL5, provided the air supply to both lagoons. Usual operating conditions required only one compressor to be operative and it was assumed that the air flow was distributed between the two aerated ponds with an estimated 80% of the airflow supplied to Lagoon No. 1 and the remainder to Lagoon No. 2. The air flow regulation and distribution was effected by the number and size of air orifices providing air directly to each Helixor. The air header piping and valving was arranged to permit the control of air to a pair of aeration units in most instances; however, several aeration headers supplied air to a single unit. The valving in the aeration headers were used only for fully open or closed conditions wherein the inlet orifices controlled the distribution of air flow.

It appeared that the recommendation of the Polcon Corporation Specifications on concrete weights for holding down the air lines and Helixors was inadequate. An inspection of the air lines and Helixors after the initial start up period was performed in the spring of 1970 by a scuba diver determining the location of the air lines, Helixors, and concrete weights in reference to the bottom of the lagoon. It was noted that not all of the Helixors were positioned on the bottom of the lagoon. Rather, some were 18" - 24" above the bottom. Approximately 500 pounds of extra concrete weights were added to each individual lateral.

KENT CHEESE WASTEWATER LAGOONS

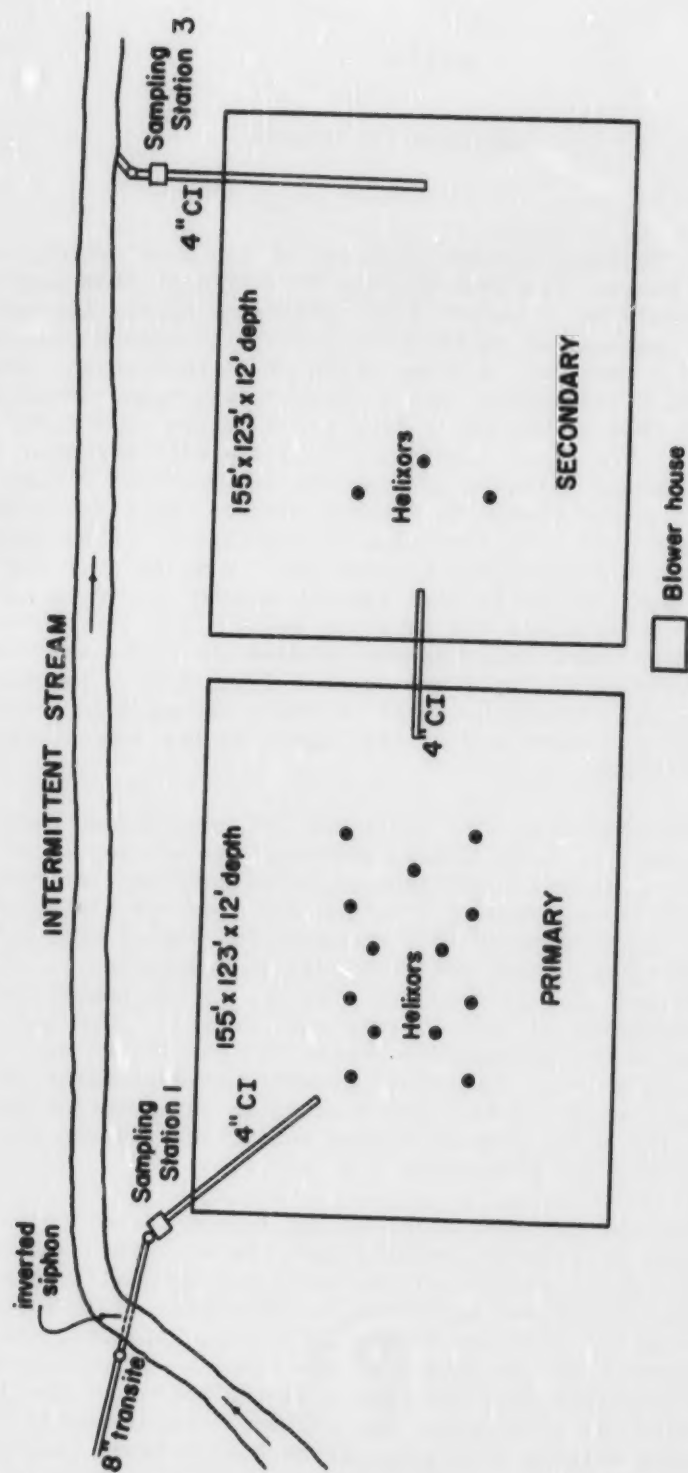


FIGURE 1

Six additional Helixors were installed as per Polcon Corporation specifications utilizing 250 pound concrete weights under each additional Helixor . These six Helixors and lines remained in their original positions.

The wastewater was conveyed to the treatment lagoons through about 3000 ft. of 8 inch transite pipe, passes through a 4 inch inverted siphon to a Snyder-Teague sampling station and thence is discharged through a 4" cast iron pipe to a point of entry to Lagoon No. 1 at approximately mid-depth. The effluent from Lagoon No. 1 passed through a submerged 4 inch cast iron pipe to the secondary lagoon. The water surface elevation in both lagoons was controlled by the placement of 4 inch cast iron riser pipe with the inlet to this pipe, or overflow from the lagoons, at a fixed elevation to maintain the 12 foot water depth. The effluent from the secondary lagoon passed through another Snyder-Teague sampling station.

The sampling stations provided flow measurements and flow composited samples for the raw waste water and treated effluents. Serious operating difficulties resulted in the raw waste sampling station due to the accumulation of cheese solids upstream from the flow measuring control section. The sampling device operated on the basis of taking a fixed volume aliquot on the discharge side of the flow measuring weir. The sampling device was activated by the water depth on the upstream side of the flow measuring weir; consequently, when the upstream float was supported by accumulated solids in the absence of influent flow, the sampling device was activated but no wastewater aliquot could be collected for the composite sample at these times. Thus, the composited sample for quality determination was considered to be reliable, whereas the flow measurements required correction.

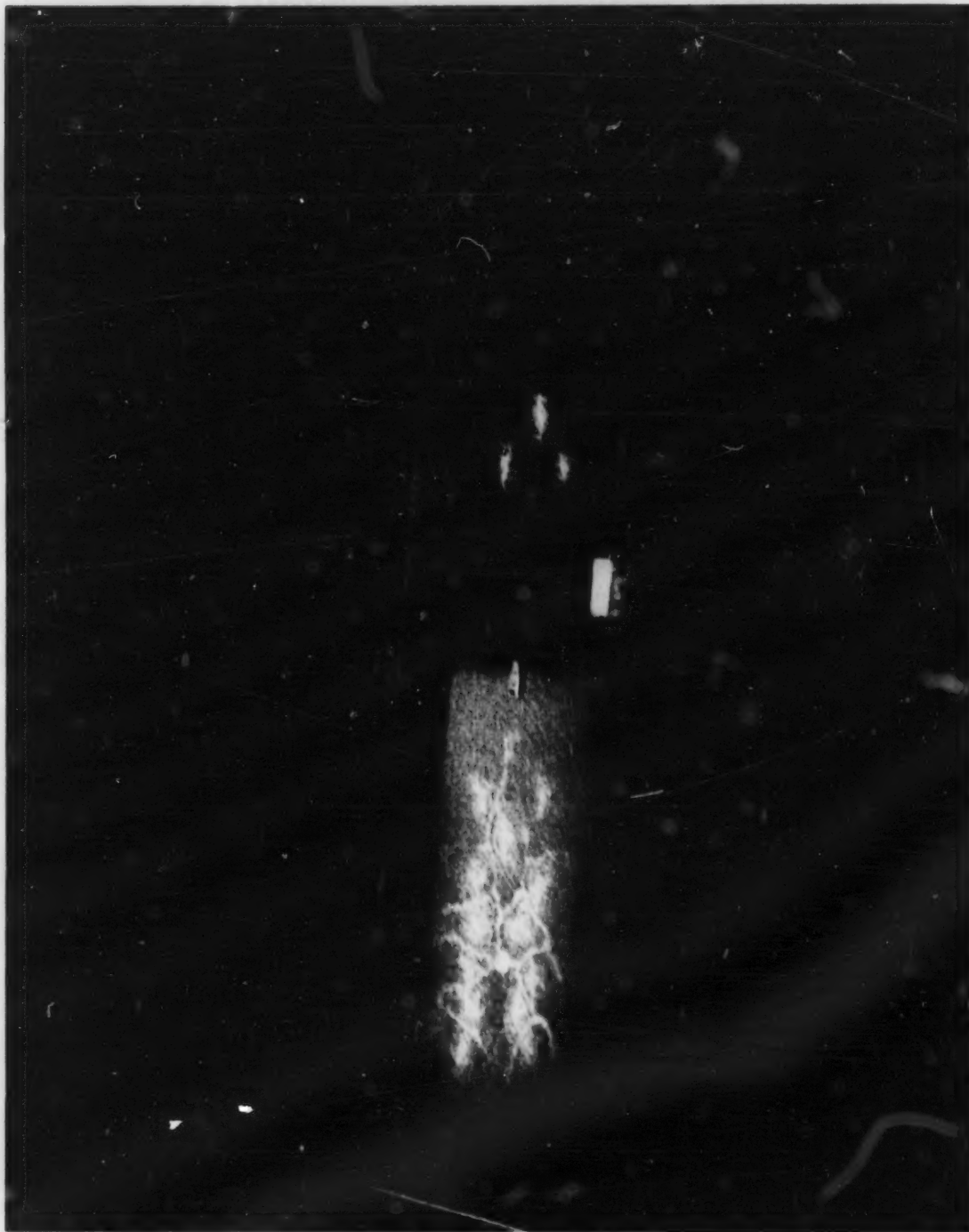
The accumulation of solids resulted in inaccurate flow measurements. The control section was modified somewhat but proved to present difficulties throughout the period of the study. In order to obtain reliable measurements of flow, it was necessary to meter water use at the cheese processing plant and to determine boiler feedwater requirements and the extent of infiltration of the conveying sewerage system for reliable flow information. Routine measurements of water pumped at the cheese processing plant were made throughout the post construction study. The well pumps were calibrated and clocks were employed to record total water used. Boiler feed water was also monitored routinely. Only after considerable effort had been exerted to correct the installed flowmeters at the lagoon site was it determined that corrected water use data would have to be employed for flow estimates to the lagoons. At that time infiltration studies were conducted over two weekends, Dec. 4-5 and Dec. 18-19, 1971, to provide an estimate of this flow contribution. During these two periods infiltration rates of approximately 690 and 900 gal/d were measured. The pipeline conveying the cheese process waters to the lagoon system was constructed in 1969 and employed a 12 inch diameter Armco truss pipe with chemical band joint. The manufacturer estimates a maximum infiltration rate with this type of construction

of 100 gal/inch diameter/day. This would result in a value of 875 gal/day based on 3,860 ft of 12 inch pipe. An average of 800 gal/d correction was employed to all water use data recognizing that this value would undoubtedly fluctuate with season. The corrections that were employed are tabulated in Appendix C.

A sampling station was not provided between the primary and secondary lagoons; thus, grab samples of the primary lagoon contents were taken near the effluent discharge pipe. Because of the long detention times experienced in each lagoon, 45-75 days, this procedure was deemed acceptable.

An aerial photograph of the wastewater treatment system appears in Figure 2.

FIGURE 2
AERIAL PHOTOGRAPH OF LAGOON SYSTEM



SECTION V

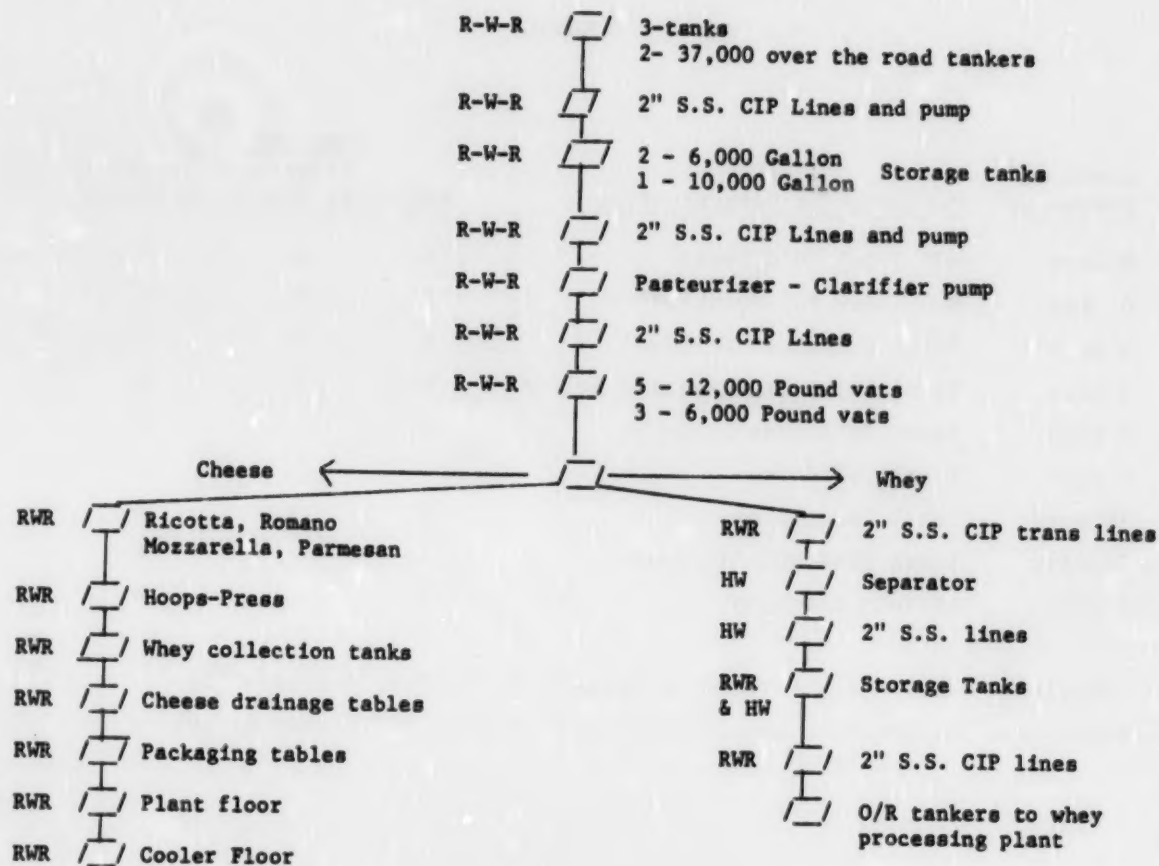
PROCEDURES

The treatment system handled the wastewater from a cheese making industry specializing principally in products of Ricotta, Parmesan, Romano, and Mozzarella cheese. The sources of wastewater were primarily from rinses and washes associated with the storage of milk, transmission lines, vats and pasteurizer with a limited amount of wastes of domestic origin. The by-product whey was collected and transported to another site for recovery and only whey wastes associated with washing and rinsing were discharged to this treatment system. The sources of wastewater treated by the aerated lagoon system are shown in Figure 3. The quantity or quality of rinsewater from each operation was not determined in this study, but rather the collective properties of the total discharge to the lagoon system were determined.

Certain measurements were made daily for operation of the treatment works such as D.O., pH, alkalinity, settleable solids, temperature, and flow quantity whereas more detailed analyses for evaluating the performance of the treatment system were obtained for raw wastewater influent, primary lagoon contents near effluent structure, and secondary lagoon effluent on an eight day sampling frequency. This permitted each day of the weekly operation to be sampled every 56 days throughout the one year study commencing on January 14, 1971. A sampling schedule is outlined in Table 1. All the analyses indicated in Table 1 were performed by Corning Laboratories, Inc., Cedar Falls, Iowa. After sample collection, appropriate volumes of well mixed sample were placed in separate sample bottles for coliform determination, for nitrogen and phosphorus analysis, and for BOD and solids analysis. Nitrogen and phosphorus samples were preserved with mercuric chlorides as prescribed by Methods of Chemical Analyses (1). Samples were shipped immediately after collection to the Corning Laboratories, Inc. where analyses were initiated no more than 10 to 12 hours after collection. The lagoon system had been in operation approximately eight months prior to initiating the sampling program; thus, the data presented does not represent start-up conditions, but there were indications that the performance did not reach steady state in all respects particularly regarding seasonal variations.

FIGURE 3

SOURCES OF CHEESE PLANT WASTEWATER



RWR = Rinse - Wash - Rinse

HW = Hand Wash

CIP = Cleaned - in - Place

O/R = Over the road tanker

Water Supply = 1/2" diameter hose
60 psi Pressure

Hot Water = Supply provided by steam/water
mixers - air flow 25 GPM

[] Domestic Waste
20 employees

[] Company Residents

TABLE 1
SAMPLING SCHEDULE*

Sampling Frequency	Parameter or Determination	Sampling Points		
		Raw	Primary Effluent	Secondary Effluent
8 days	BOD 5 day 20°C Total	X	X	X
8 days	BOD 5 day 20° Soluble**	X	X	X
8 days	Total Suspended Solids	X	X	X
8 days	Total Volatile Suspended Solids	X	X	X
8 days	Total Coliforms	X		X
8 days	Fecal Coliforms	X		X
30 days	Total Phosphorus	X		X
30 days	Total Kjeldahl Nitrogen	X		X
30 days	Nitrate Nitrogen	X		X

*Analyses performed according to Methods of Chemical Analysis (1)

**Analysis performed on sample filtrates (Whatman #42 filter paper)

Additional operating and performance evaluations were performed during the experimental period including the following items:

1. D.O. measurements with respect to horizontal and vertical control in the primary lagoon,
2. measurements of velocity in the horizontal plane at various locations and depths within the primary lagoon,
3. measurement of oxygen uptake of the treatment lagoon contents,
4. determination of the oxygen transfer coefficients of α and β to enable the evaluation of the aeration system employed.

Various other operating conditions and observations were recorded on a routine basis to assist in the overall evaluation of the performance of this treatment system (Appendix C).

Special measurement techniques employed for the purpose of evaluating the aeration system employed a grid established in the primary lagoon above the water surface to provide a measuring base for horizontal control. At the time oxygen transfer was evaluated, it was necessary to have a measureable D.O. in the lagoon and evidence that the D.O. was maintained at a uniform or steady state condition. The oxygen uptake or demand of the lagoon contents were determined by taking a number of representative samples of the mixture under aeration and placing them in bottles capable of excluding further oxygen transfer from the atmosphere and wherein the decrease in D.O. concentration was measured with respect to time with a Yellow Springs Instrument D.O. probe. In order to evaluate the influence of photosynthetic plankton on the oxygen transfer in the lagoon, oxygen uptake measurements were made both on rates observed under light and dark conditions. There was no discernible difference in the rates observed which indicated that oxygen transfer or supply from this source was negligible during the test periods.

Aerator efficiency was determined on the basis of line to water for the complete aeration system wherein power input was metered in each instance for amperage and voltage drawn. The results were presented in terms of pounds oxygen transferred per kw-hr. Thus, the reported efficiency included oxygen transferred from the atmosphere as well as oxygen transferred by virtue of the aeration system employed.

Aeration efficiencies were corrected to standard conditions of zero dissolved oxygen and 20°C with appropriate corrections for α and β . Representative samples of the lagoon contents were tested in Madison laboratories in a simulated diffused air system to compare transfer capacities (α) for the waste mixture against tap water for two replicate samples. The value α was obtained as an average of the ratio of the transfer rates in the wastewater and tap water. The value β was

determined on the basis of the highest D.O. obtained in the aerated lagoon contents for a given temperature, checked by the Winkler D.O. method, against the D.O. at saturation for tap water at the same temperature.

The velocity profiles in the aerated lagoon were made with the use of a Gurley Current Meter which was fixed in the horizontal plane by attachment to a vertical rod. The current meter was rotated in the horizontal plane at a predetermined depth and the maximum velocity and direction were noted to obtain a vectorial representation of the water movement in the lagoon. The velocity was observed in four directions parallel to the sides of the lagoon in the central portion of the aerated lagoon where a single maximum velocity component was not observed due to the highly undirected flow patterns evident within the area bounded by the aeration devices. An attempt was made to measure the vertical velocity component immediately above the discharge of the aeration device but the variation in fluid density as a result of high levels of air entrainment caused on the measurements to be somewhat erratic. Measurements were made one foot below the surface, mid-depth and one foot from the bottom.

SECTION VI

RESULTS AND DISCUSSION

Aerator Performance

The performance of the Polcon Corporation subsurface aerators (Helixors) was evaluated over the one year period in the primary lagoon only. As indicated earlier, the lagoon was sampled at a number of points on a grid (Figure 4) at selected depths to determine the oxygen uptake rates, dissolved oxygen concentration, and temperature. Tests were repeated five times over the grant period in order to evaluate performance under different waste loading and environmental conditions.

Oxygen Uptake Rates - The average oxygen uptake rates recorded in the primary lagoon over the one year study period are presented in Table 2. The values reported represent the average uptake measured at a number of selected points within the lagoon (see Appendix A). Normally, the uptake rates measured on any given day were within 10 percent of each other. The uptake rates reported were not converted to specific uptake rates ($\text{mg O}_2/\text{gVSS}/\text{day}$) since the volatile solids fraction measured in this lagoon was not well correlated with the active biomass during the study period. Large amounts of suspended volatile matter non-biological in origin associated with the wastewater contributed significantly to the total volatile fraction measured. The influence of algal cells on the oxygen uptake rates in the primary lagoon were negligible as measured by both dark and light bottle uptake rates.

Examination of the uptake rates presented in Table 2 suggest that the biological activity in the primary lagoon was comparable to other aerated lagoon systems (2). The analysis obtained in May, 1971, occurred during the anaerobic period and indicates that the lagoon was extremely active. As discussed later, this high rate of oxygen consumption was believed to be due to the rapid solubilization of organic matter and synthesized cells which had settled in the lagoon and had subsequently been stored over the cold winter months.

Oxygen Transfer Capacity - The values of the oxygen transfer capacity of the primary lagoon contents, as measured by the ratio of the oxygen transfer coefficients in the waste to that in tap water, are reported in Table 3 as alpha. These values represent the average values obtained from several points within the lagoon. Alpha appeared to fluctuate seasonally being highest during the summer months. There was not sufficient data to establish any reliable correlations between alpha and lagoon BOD solids or temperature, but it is reasonable to assume that changes in the metabolic activity within the lagoon would result in changes in the oxygen transfer relationships.

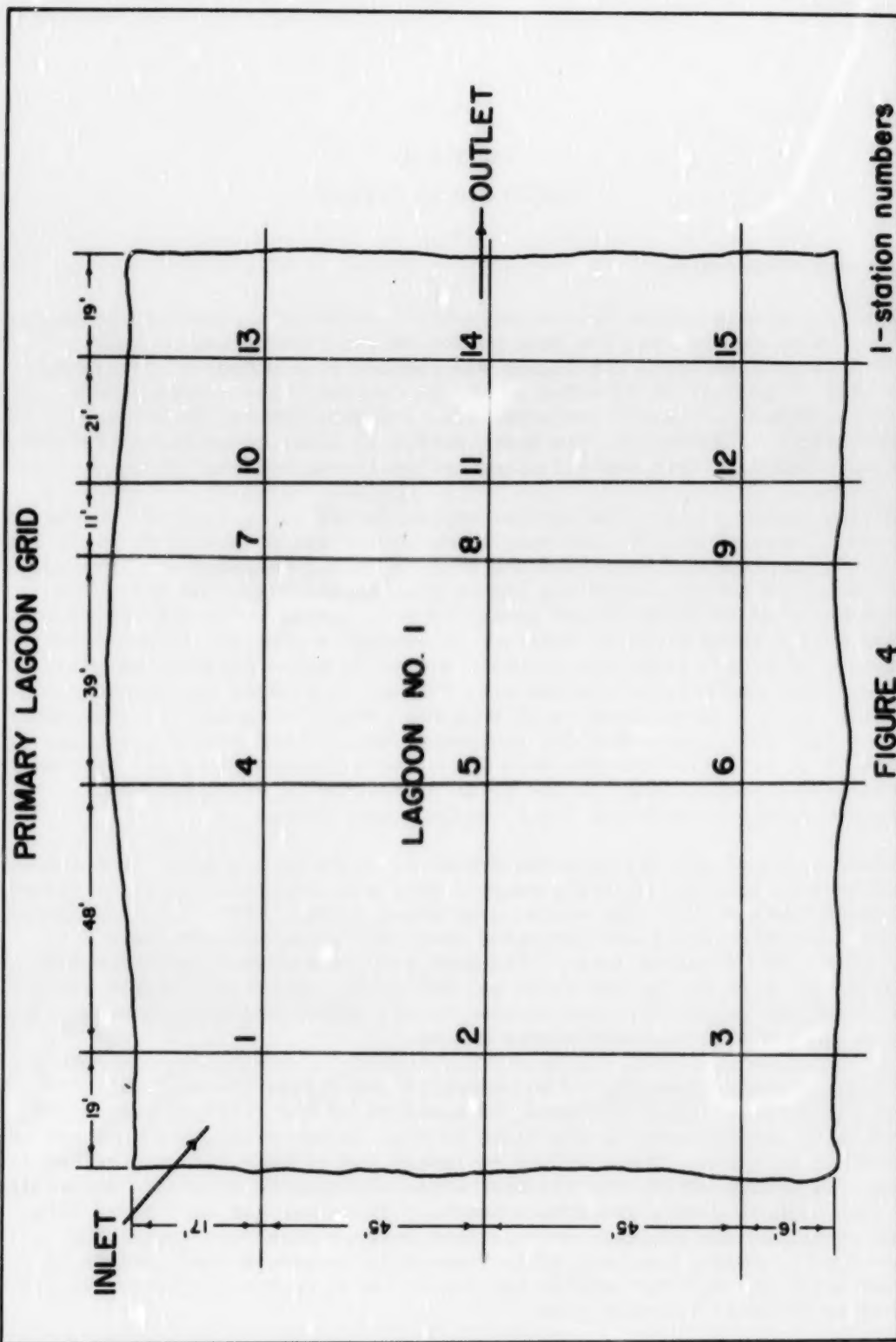


TABLE 2
OXYGEN UPTAKE RATES
PRIMARY LAGOON
(Average Values)

DATE	TEMP. °C	D.O. mg/l	UPTAKE RATE	
			mg/l/hr	lb/hr
11-19-70	6.3	2.8(43)	2.1 (5)	16.7
5-23-71	17.0	0.0(43)	5.6 (5)	44.6
7-21-71	24.0	4.4(39)	1.2 (4)	9.55
8-26-71	23.5	4.4(27)	0.8 (4)	6.4
10-7-71	16.5	2.3(41)	1.5 (3)	12.0

() - Number of samples analyzed (see Appendix A & Table 4)

TABLE 3
OXYGEN TRANSFER DATA
PRIMARY LAGOON

DATE	TEMP. °C	D.O. mg/l	α	β	STD. TRANSFER ¹	
					lb/kw-hr	lb/hp-hr
11-19-70	6.3	2.8	0.75	0.96	5.25	3.91
5-23-71	17.0	0.0	0.88	0.96	----	----
7-21-71	24.0	4.4	0.90	0.98	3.18	2.47
8-26-71	23.5	4.4	0.94	0.98	0.98	0.73
10-7-71	16.5	2.3	0.70	0.98	3.45	2.59
Overall Average ²			0.81	0.97	3.96	2.98

1 - 20°C, D.O. = 0.0, Tap water - see Appendix B

2 - Excluding 8-26-71 data for reason on page 19.

The solubility of oxygen was only slightly affected by the wastewater characteristics within the primary lagoon. Normally values of beta in long detention-type processes such as the one studied here approach 1.0.

Aeration Efficiency - The aeration efficiency of the Polcon Corporation Helixors are presented in Table 3 for five test days. The aeration efficiencies were computed by employing the measured oxygen uptake rates, dissolved oxygen concentration, alpha, beta, and temperature. Power was metered during the test period. The value of the aeration efficiency was corrected to standard conditions of 20°C, tap water (alpha and beta equal 1.0), and a dissolved oxygen concentration of 0.0 mg/l. The approximate saturation value for oxygen for the 12 foot lagoon depth was estimated by using the relationship given by Oldshue(3):

$$C_s = C \left\{ \frac{P_b}{20.4} + \frac{O_t}{42} \right\} \quad (1)$$

where C_s is the oxygen saturation value corresponding to the average partial^s pressure of oxygen in the gas stream entering and leaving the aerator, C is the oxygen saturation value of water at atmospheric pressure, P_b is the absolute pressure in psi at depth, d and O_t is the percent concentration of oxygen in the air leaving the lagoon. A sample calculation appears in Appendix B.

The overall aeration efficiencies in the primary lagoon were estimated by assuming that 80% of the air provided by the blower was directed to the primary lagoon, since 13 of the 16 helixors were located in the primary lagoon. During August 26, 1971, test, both blowers were in operation resulting in a doubling of the power input to the aerators. Yet the actual measureable oxygen transfer rate was not appreciably higher than for other test periods. It has been well documented in the literature (4,5) that aeration efficiency decreases with increased air flow rates and this analysis tends to verify that fact. In addition, however, it was determined that, during the operation of both compressors, the line pressures increased very significantly requiring the bleeding of air from one of the headers. This was practiced by the use of a "blow-off" or "by-pass" Helixor in the secondary lagoon. Thus, the poor transfer rates for August 26 were not included in the overall average for aeration efficiency.

The measured transfer efficiencies of the Helixor units were lower than the expected performance claimed by the manufacturer (6). Whether this was due to undersizing of the air headers, the geometry of the Helixor-basin configuration, or was, in fact, the actual field expectation for these units has not been determined but it suggests that caution must be exercised in sizing aerators for lagoon systems. An additional factor to be considered was the oxygen uptake occurring in the benthal deposit. No effort was made to ascertain what this contribution was, but significant uptake from this source would substantially increase the aeration efficiencies reported herein.

Oxygen Dispersion - An important function of aeration devices in biological processes is the dispersion of oxygen throughout the basin contents. Normally, rule of thumb criteria recommend that at least 6 to 10 horsepower per million gallons (hp/MG) be supplied to insure adequate oxygen dispersion (7). For the primary lagoon, 8.5 hp/MG were provided during the single blower operation. The results of dissolved oxygen analyses in the primary lagoon for the five field surveys are presented in Table 4. Examination of this data indicates that oxygen was well dispersed throughout the primary lagoon.

Mixing and Solids Suspension - The D.O. parameter is indicative of the dispersion of substances in the dissolved state such as soluble B.O.D. However, the distribution of particulate matter such as biological floc would be a function of the velocity distribution and the velocities necessary to keep the particulate matter in suspension. Normally in aeration tanks in wastewater treatment, velocities along the bottoms of these units would range from 1.0 to 2.0 fps depending upon the geometry of the aeration system employed. Velocity measurements within the lagoon can serve only to show flow regimes developed within the lagoon with the given aeration equipment for the placement arrangements employed and the geometry of the lagoon. Effectiveness of the system in terms of keeping particulate solids in suspension can be evaluated in a qualitative way by determining the location and extent of solids deposition on bottom surfaces of the lagoon.

The maximum velocities observed in the primary lagoon at the various depths are reported in Table 5 for one blower, 192 cfm, and two blowers, 384 cfm. These airflow rates correspond to 2.3 cfm per lineal foot and 4.6 cfm per lineal foot respectively along the longest bottom dimension of the lagoon. The vectorial representations of the resulting surface velocities are shown in Figures 5 and 6 for the single and dual blower operation.

An estimated surface wind of less than 5 mph as indicated on the diagram may have influenced the resulting velocity measurements. Because of the upward vertical velocity components in the immediate vicinity of the inlet and outlet of the Helixors and the necessary compensating downward movement of water between adjacent aeration units, no definitive flow pattern is discernible. However, in zones peripheral to the aeration section, where the bottom of the lagoon is sloped to intersect the water surface, horizontal velocities were measured near the water surface indicating a general circulation pattern towards the periphery of the aerated lagoon with velocities ranging from less than 0.1 to 0.8 fps. Likewise, to compensate for the outward movement of water toward the periphery of the lagoon near the surface, water movement must be in the opposite direction near the bottom of the lagoon, the measurements and magnitude of which was less discernible. Thus, a circulatory pattern of flow developed, in the outer prism shaped sections similar to that which is depicted in Figure 7.

TABLE 4
DISSOLVED OXYGEN CONCENTRATIONS*
PRIMARY LAGOON

Date	1	2	3	4	5	6	7	8	9	10	11	12	13	14	15
11/19/70	2.7	2.9	2.9	2.7	3.1	2.9	3.0	2.8	2.8	2.8	2.8	2.7	2.8	2.7	2.6
1 blower	2.7	2.8	2.9	2.7	3.0	2.9	2.9	2.8	2.8	2.8	2.7	2.7	2.7	2.7	2.5
	--	2.8	2.7	2.7	3.0	2.9	2.9	2.8	2.7	2.8	2.7	2.7	2.6	2.7	--
5/14/71	0.1	0.1	0	0	0.3	0.1	0	0.3	0.1	0.1	0.3	0	0	0.1	0.1
1 blower	0.1	0.1	0	0	0.2	0.1	0	0.2	0.1	0.1	0.3	0	0	0.1	0
	--	0.1	0	0	0.2	0	0	0.1	0	--	0.3	0	0	0.1	0
7/21/71	4.4	4.4	4.4	4.3	4.3	4.4	4.6	4.7	4.2	4.8	4.4	4.3	4.9	4.7	4.4
1 blower	4.3	4.2	4.3	4.3	4.3	4.3	4.7	4.7	4.2	4.7	4.4	4.2	4.8	4.7	4.4
	--	4.2	--	3.0	4.3	4.2	4.7	4.5	4.2	4.7	4.3	--	--	--	--
8/26/71	4.1	4.3	4.7	4.3	4.4	4.7	--	4.7	--	--	4.4	--	4.4	4.5	4.5
2 blowers	3.9	4.6	--	4.3	4.4	4.7	--	4.6	--	--	4.4	--	4.4	4.5	--
	--	3.0	--	4.4	4.5	4.7	--	4.6	--	--	4.5	--	--	4.5	--
10/7/71	2.2	2.2	2.7	2.3	2.3	2.4	2.3	2.4	2.3	2.3	2.1	2.3	2.4	3.0	--
1 blower	2.0	2.1	2.7	2.4	2.3	2.4	2.3	2.2	2.3	2.3	2.1	2.3	2.3	2.6	--
	2.1	2.0	--	2.3	2.3	2.4	2.3	2.2	2.3	2.3	2.1	2.3	2.3	2.5	--

*D.O. values in mg/l at 1 ft

3 ft

5 ft below water surface

TABLE 5
VELOCITY DISTRIBUTIONS IN PRIMARY LAGOON
November 7, 1971

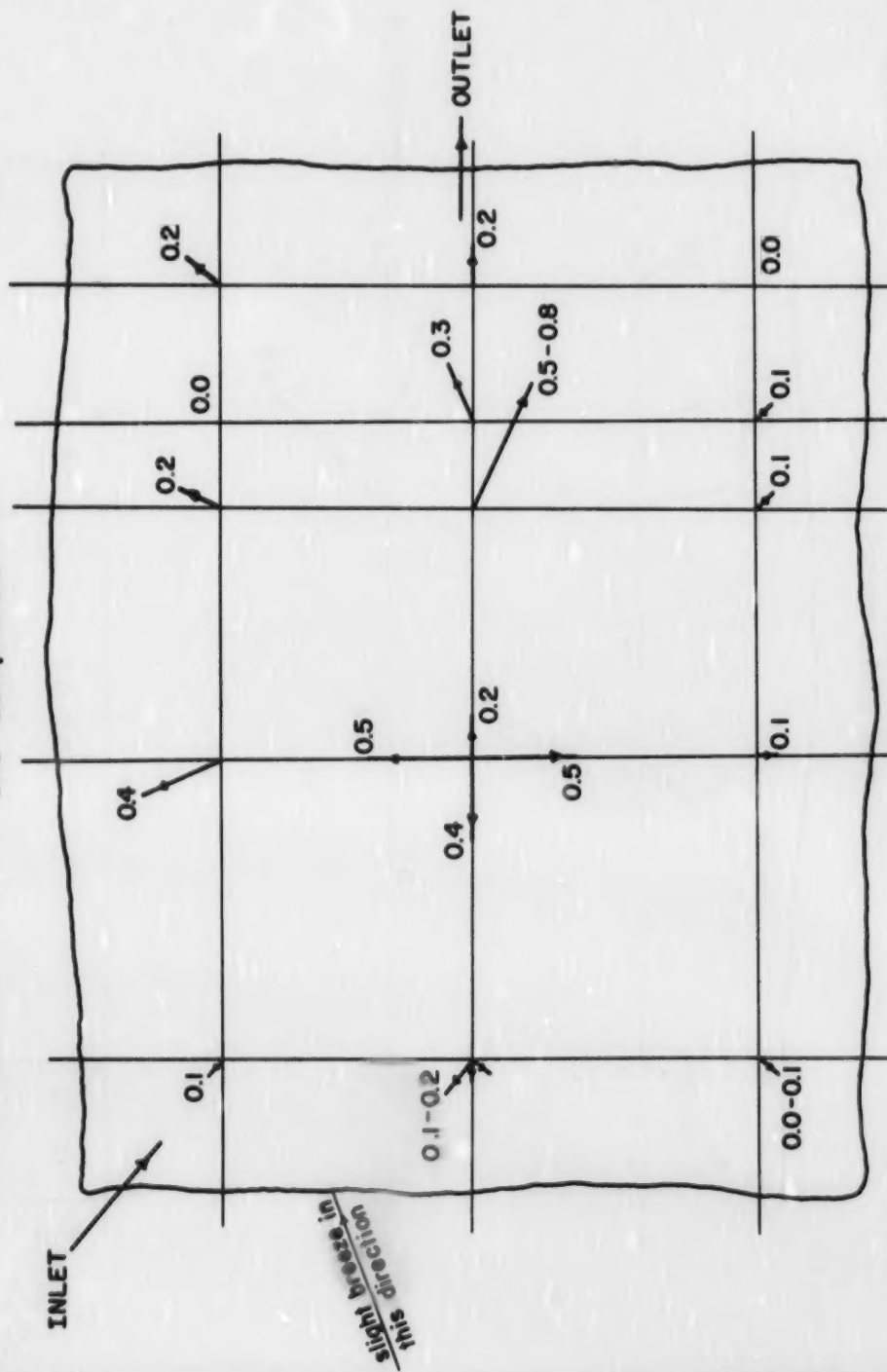
Station	One Compressor			Two Compressors		
	Surface*	Mid Depth	Bottom**	Surface*	Mid Depth	Bottom**
1	0.1	0.0	0.0	0.0-0.2	0.0-0.2	0.0
2	0.1-0.2	0.0	0.0	0.1-0.2	0.0-0.1	0.1-0.2
3	0.0-0.1	0.0-0.1	0.0	0.2-0.3	0.2-0.3	0.0-0.1
4	0.4	0.1	0.0	0.6-0.8	0.3-0.4	0.1-0.2
5	0.2-0.5	0.2-0.3	0.1	0.2-0.6	0.2-0.3	0.0-0.2
6	0.1	0.0	0.0	0.0-0.2	0.0-0.2	0.0-0.1
7	0.2	0.0	0.0	0.4-0.5	0.0-0.2	0.0-0.1
8	0.5-0.8	0.2	0.2	0.8	0.2	0.2
9	0.1	0.0	0.1	0.2	0.1	0.1
10	0.0	0.2	0.0	0.4-0.5	0.0-0.1	0.0
11	0.3	0.2	0.1	0.3	0.2	0.1-0.2
12	0.1	0.0	0.1	0.1	0.1	0.0
13	0.2	0.0	0.0	0.2	0.0	0.0
14	0.2	0.1	0.0	0.2	0.0	0.1
15	0.0	0.0	0.0	0.1	0.0	0.0

All values of velocity in ft/sec.

* One foot below

**One foot above

VELOCITY PROFILE AT ONE FOOT BELOW SURFACE one compressor



Date Nov. 7, 1971

FIGURE 5

[illegible]

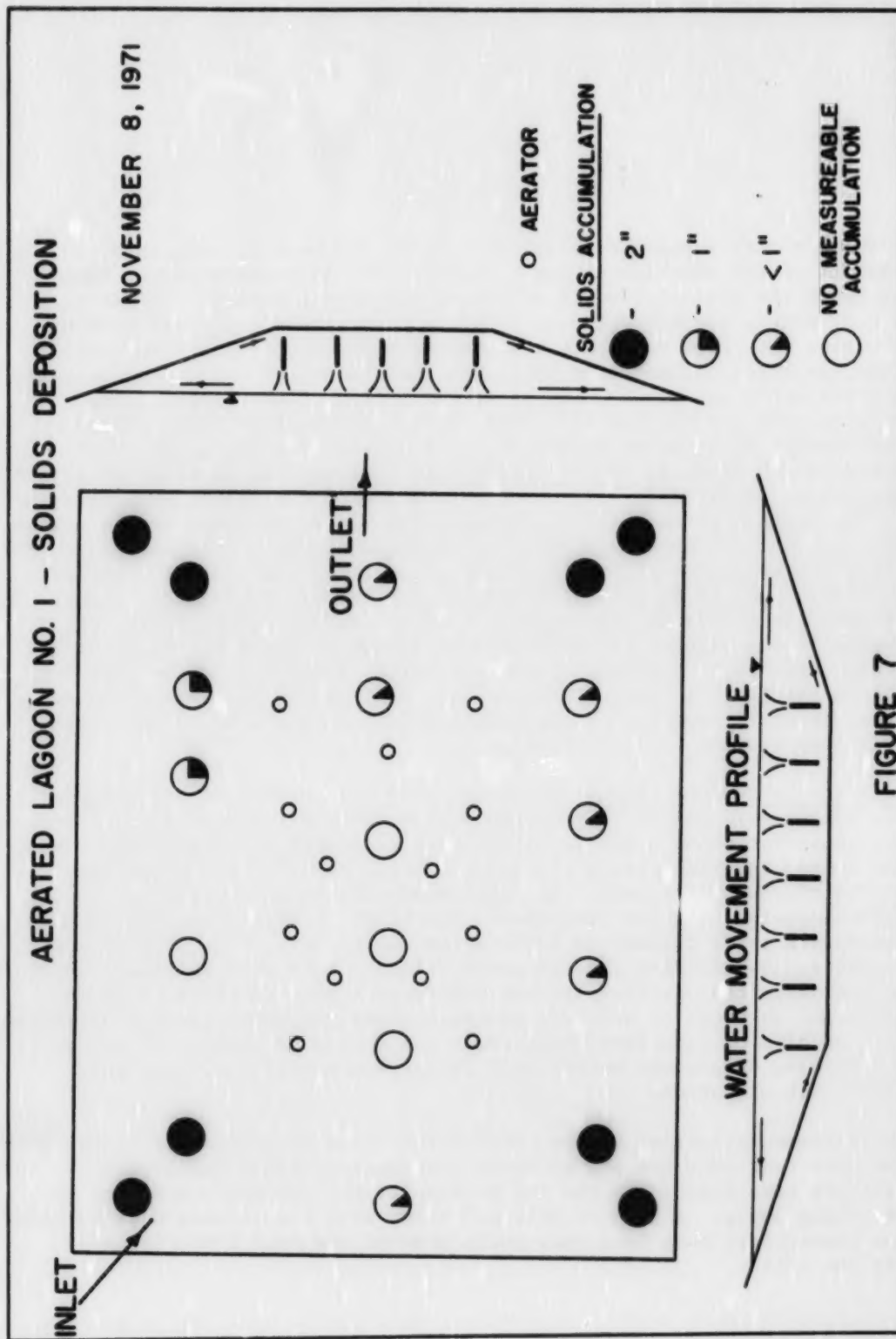


FIGURE 7

Perhaps a more meaningful observation that was made to ascertain the ability of the aeration system to keep solids in suspension was to estimate the accumulation of solids on the lagoon bottom. Utilizing a D.O. probe, a crude estimate of the sludge accumulation was made by lowering the probe to an elevation where the D.O. decreased to zero. This elevation was compared with that where a weight would come to rest, a point which was presumed to be at the bottom of the lagoon. The results were plotted qualitatively by the graphical representations for deposit as shown in Figure 7. It is noted that the greatest accumulation reported on November 8, 1971, almost two years after lagoon operation, occurred in the non-aerated zones of the lagoon corresponding to the tapered prism sections representing sloped sections of the lagoon nearest the intersection of the two adjacent sloped bottom sections. The least discernible deposits of solids occurred in the central aeration zone which may be due either to the higher levels of fluid turbulence to keep the solids in suspension or due to the greater diffusion of oxygen in the lighter accumulations of sludge at these points. For longer periods of treatment plant operation, it is likely that greater accumulations will occur. However, at the time the determinations were made, it would appear that the accumulation of solids would not present a serious problem for this system.

Aeration System Operation - Maintenance of low ambient temperature in the compressor building during the months of July and August required the installation of a fan to maintain an equilibrium temperature between the inside and the outside of the building. During the extreme cold months of the winter, when the temperature of the water drops to approximately 4 degrees Centigrade, the brake horsepower of the fifteen horsepower motor climbed to 18-19 brake horsepower. This problem was corrected by utilizing the six extra Helixors that were initially installed to discharge the air whenever two compressors were required to be in operation in order to maintain residual dissolved oxygen in the springtime. This problem of increased brake power could also be handled by replacing the fifteen horsepower motors with twenty horsepower motors in this particular situation.

The recommendations of Polcon Corporation for stainless steel clamps with the ends bent over has proven to be inadequate. Since January, 1971, three Helixors have floated to the top because of the stainless steel clamp vibrating loose. A type of bolt-nut clamp with the threads damaged after the connections have been made would prevent the nuts from vibrating off the bolts.

Lagoon Performance

The performance of the staged lagoon system was interpreted in terms of several measured parameters: BOD (total and soluble), suspended solids, nitrogen, phosphorus, coliforms (total and fecal). The results reported herein are subdivided in accordance with these parameters. Tabulated results of lagoon performance appear in Appendix C.

The loading to the lagoon system is summarized for the one year study in Figure 8. It is apparent from this figure that flow and organic load to the lagoon is quite variable over the one year period. Hourly flows over a typical 24 hour period are presented in Figure 9 for two days, January 24 and 25, 1972. Analysis of this data indicates that the ratio of maximum to average flow for the day was 1.9. The average flow rate over the 12 month study was approximately 17,000 gallons per day including Sundays.

The apparent cyclic variation in BOD and solids loading to the lagoons (Figure 8) are due to the sampling schedule employed. As noted in Table 1, samples were collected every 8 days in order to obtain data over the entire week. The flows and pollutant loads noted, therefore, represent weekly cycles. Therefore, a trend line was difficult to describe. Average BOD loading over the 12 month study was 285 lb/day and suspended solids loading was 85 lb/day. An average of 13,000 lb cheese were produced per day (excluding Sunday) requiring approximately 1.3 gallons of water per lb of cheese and resulting in 22 lbs of BOD and 6.5 lb of suspended solids per 1000 lb of cheese produced.

Biochemical Oxygen Demand - The performance of the lagoon system in removing biodegradable organic matter was measured by the five day BOD determination. Both total and filtered analyses were performed in order to reflect the influence of the pond system on actual biochemical stabilization as compared with physical separation of particulate organic matter. No long-term ultimate BOD analyses were performed nor were COD or Total Carbon Analyses; thus, the interpretation of the data in terms of mass balances of oxygen demanding materials through the system was limited.

Although the BOD load to the lagoons varied widely over the one year period of study, the most apparent influence on lagoon performance was temperature. In Figure 10, BOD load and temperature are plotted along with the primary lagoon dissolved oxygen (D.O.) concentrations. It is immediately apparent that D.O. rapidly disappeared in the spring along with rising temperatures. In fact, D.O. reached 0.0 on April 10 and did not reappear until July 19. During the remainder of July, six of thirteen days the lagoon was devoid of oxygen. Also, 11 of 31 days in August, 15 of 30 days in September, 19 of 31 days in October, 6 of 30 days in November, and 2 of 31 days in December D.O. values were less than 1.0 mg/l.

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

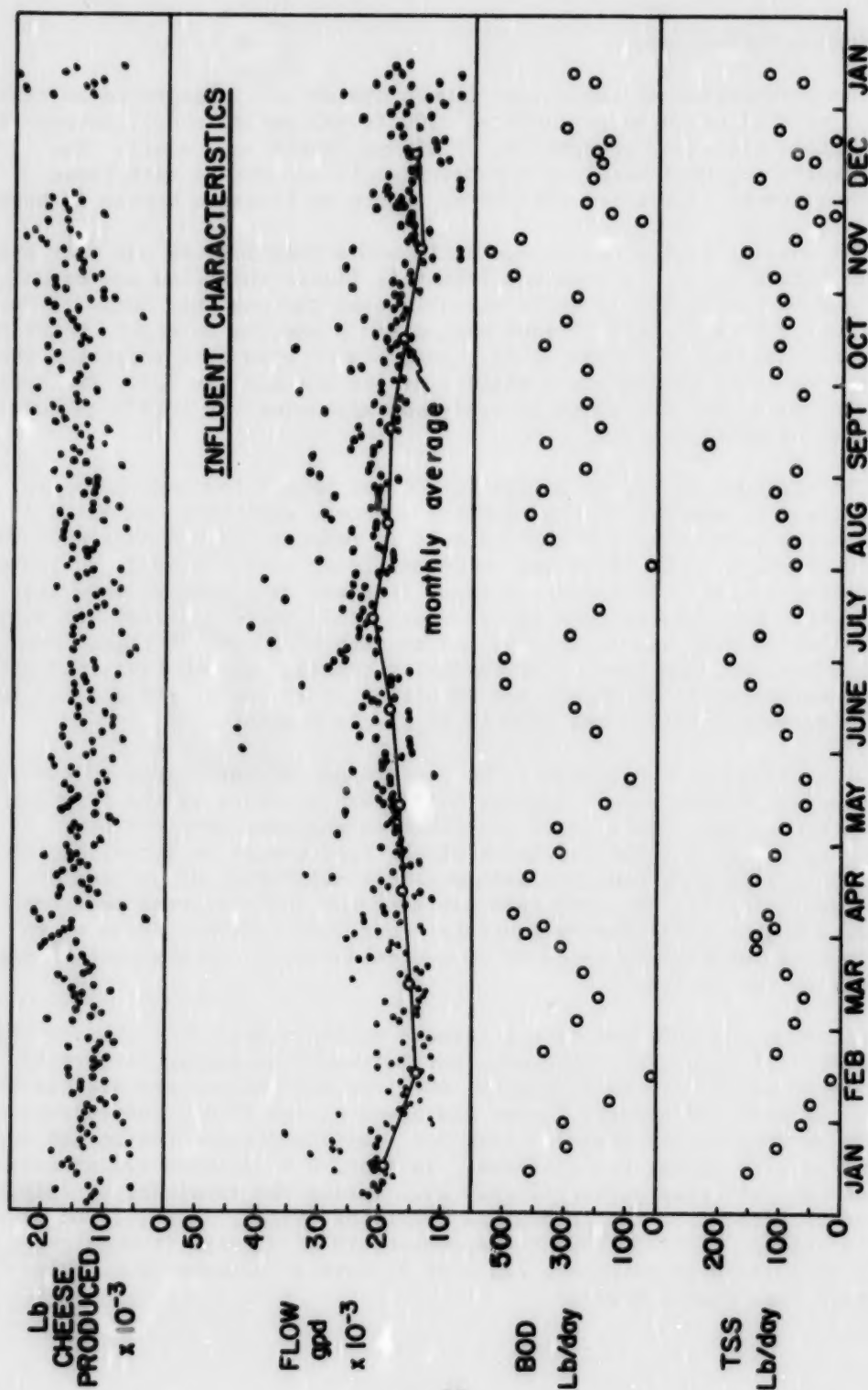


FIGURE 8

CHEESE PROCESS WASTEWATER FLOWS

JAN 1972

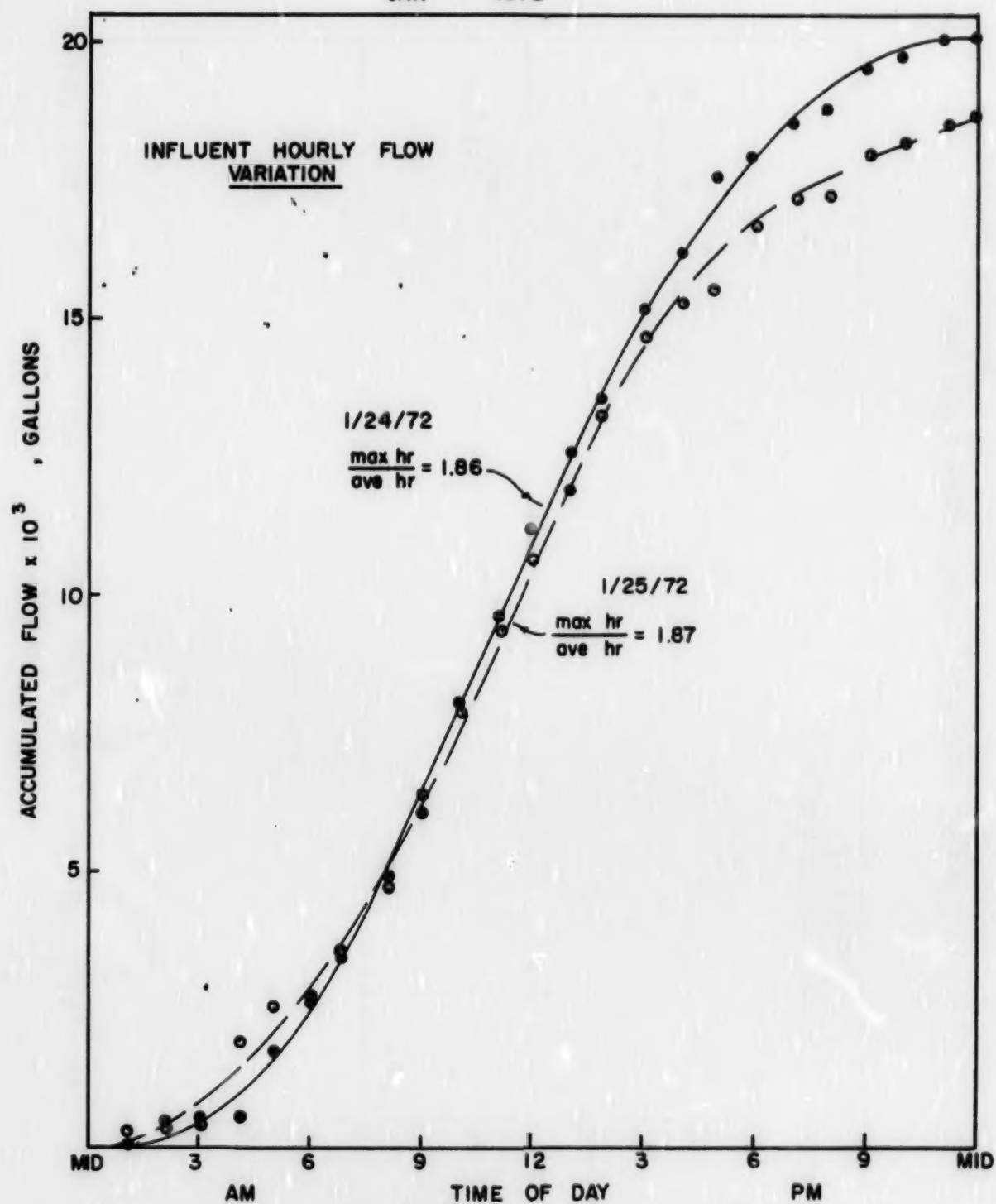


FIGURE 9

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

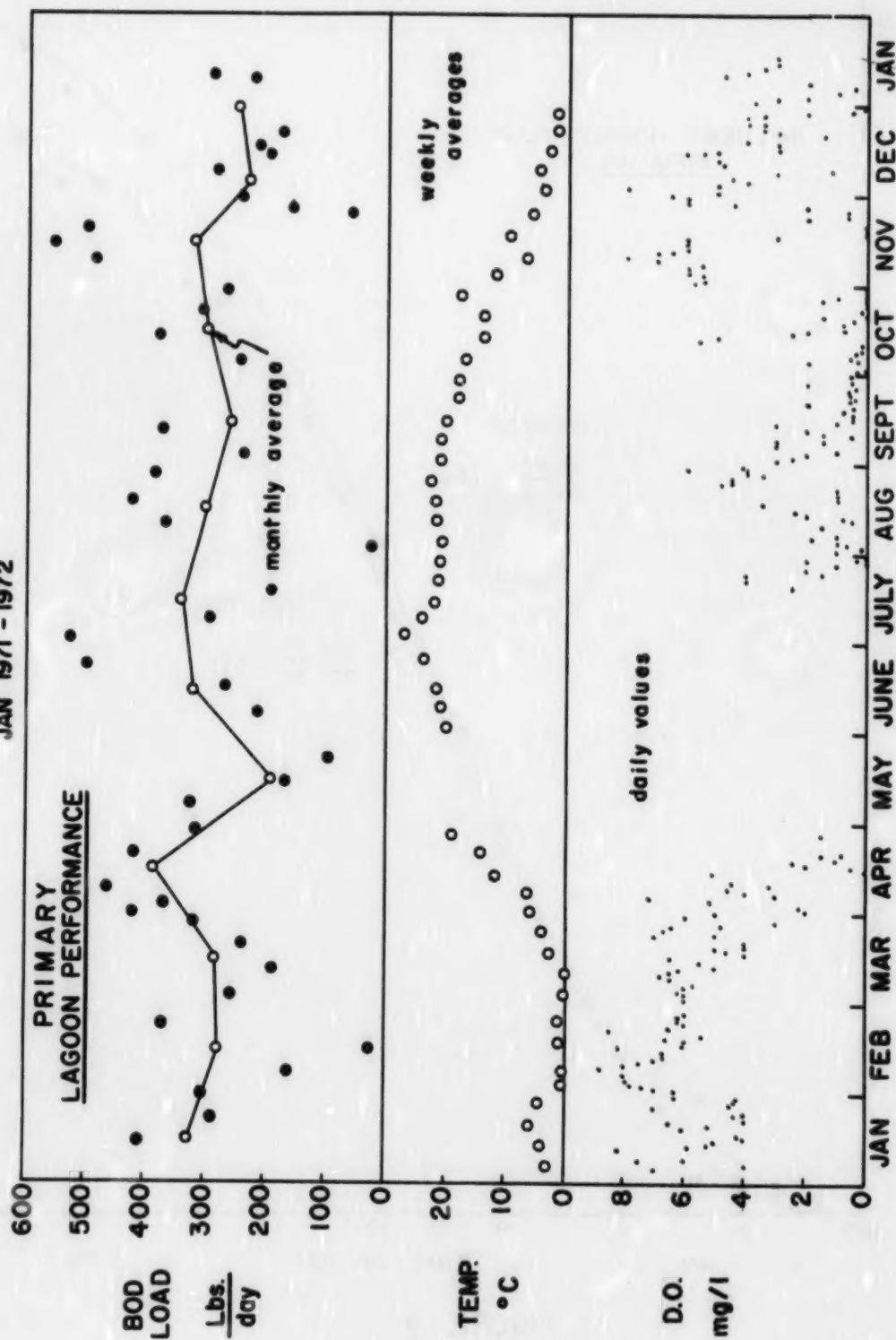


FIGURE 10

It might be assumed that this increased demand for oxygen during the spring was due to the increased biological activities brought about by the warmer temperatures. However, examination of the BOD load just preceding this period would indicate that the demands measured during this period could not be accounted for by the applied load.

For example, if one assumes that the ultimate oxygen demand is 1.47 times the 5 day value (deoxygenation constant, $k_1 = 0.1/\text{day}$), the oxygen demand for a 5 day BOD load of 285 lb/day¹ would be 419 lb/day. The oxygen demand measured in May, during the oxygen deficient period, was 1165 lbs/day, 746 lbs in excess of that predicted above. Furthermore, if one computes the oxygen supplied during this period based on a standard transfer rate of 4.0 lbs O₂ per Kw hr., approximately 600 lbs of oxygen are transferred per day by this system at 20°C and zero D.O.

Therefore, it is reasonable to assume that significant amounts of biodegradable organics are being solubilized from the benthal deposits. During the cold months, anaerobic decomposition of the settled organic matter slows down resulting in accumulations of this material. As the temperature increases, biological activity also increases and solubilization of the benthal deposits begin to appreciably contribute to the organic load of the lagoon. This solubilization increases to a maximum value and then remains constant over the summer months. It may be assumed that this cyclic effect will continue, stabilizing the settled organic solids during the warmer periods and continue to produce both gas and soluble organic compounds.

The high oxygen demands occurring during the spring in waste stabilization lagoons has been attributed to benthal digestion by a number of investigators. Marais and Capri (8) have attempted to estimate the benthal demand in aerated lagoons. In a dynamic simulation of data collected for domestic wastewater lagoons, they showed that the rate of benthal decomposition was highly temperature dependent with an estimated value of the temperature coefficient theta of 1.35. They found that approximately 40 percent of the settled organic matter was resolubilized for the domestic waste.

The one year period of this study was insufficient to adequately quantify the benthal contribution to the lagoon organic load. The study period was too brief to evaluate the effectiveness of the anaerobic digestion in reducing accumulated sludge deposits. Steady state (or rather a quasi-steady state) may occur only after a number of years. Yet, based on this and other studies (2,8,9), it is reasonable to report that oxygen demands during the spring may exceed three times the applied BOD load.

Dissolved oxygen concentrations in the secondary lagoon were normally greater than 2.0 mg/l even during the April-July period of anaerobiosis in the primary lagoon. Only during two weeks in late June did dissolved oxygen levels drop below 2.0 mg/l and there was never a time when a value of zero D.O. was recorded. Dissolved oxygen concentrations were highest during the winter months, decreasing as oxygen uptakes increased in the spring and summer months.

The results of BOD analyses in the two lagoons appear in Figure 11 and average BOD-analyses by quarter are presented in Table 6. The removal of BOD was strongly correlated with lagoon temperatures and concomitant biological activity reaching a maximum late in the summer and early fall. Average BOD removal through both lagoons was 97.3 percent with the poorest performance occurring in the winter quarter with a value of 94.5 percent. During the 4 1/2 month period, July 1 through November 15, the effluent soluble BOD did not exceed 30 mg/l and was ≤ 20 mg/l 70 percent of the time. Examination of the probability plots for total and soluble BOD in the primary and secondary lagoon effluents (Figures 12 and 13) indicates that the distributions of effluent BOD are skewed, being bounded at the lower end by the presence of organic matter relatively resistant to biological degradation. Considerable scatter of effluent BOD is demonstrated above the 50 percentile suggesting that, even with very long detention times, the reliability of the system to produce effluent BOD concentration less than 25 mg/l at all times is very improbable.

It is of interest to note that the soluble fraction of the effluent BOD represents on an average 50 percent of the total, with the highest values occurring during the spring and summer quarters.

Lagoon Modelling for BOD Removal - The mathematical modelling of biological wastewater treatment systems has become a popular method for compacting design data into a simple and useable form. Unfortunately, the biochemical, chemical, and physical reactions taking place in biological waste treatment processes are only poorly understood and the techniques of mathematical modelling normally become curve fitting processes (10). Most popular models currently in use for biological systems include the Monod equation (11, 12, 13, 14) and first order reaction models (15, 16).

Defining the appropriate model parameters for these mathematical models requires careful experimental design in order to examine the model over a rather broad range of the independent variables. In addition, the modeler must be aware of all independent variables which will influence the outcome of the reaction.

The study conducted and reported herein was not designed to provide this kind of information. The lagoons were designed to accept the flow and pollutant load that occurred and little or no control was available. Furthermore, funds were inadequate to make meaningful measurements of both dependent and independent variables. Therefore, it would seem inappropriate to try to develop any type of sophisticated model for this system. Furthermore, until one is able to define with considerable confidence, the independent variables which will provide a model with sufficient rigor to extrapolate process performance from one system to another, there is considerable hazard in presenting any type of mathematical expression which might be used in an effort to describe another system. Thus, by use of dynamic modelling, one may employ the data appearing in Appendix C to write a model for the experimental system described herein. Yet, use of this model to describe performance of any other similar

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

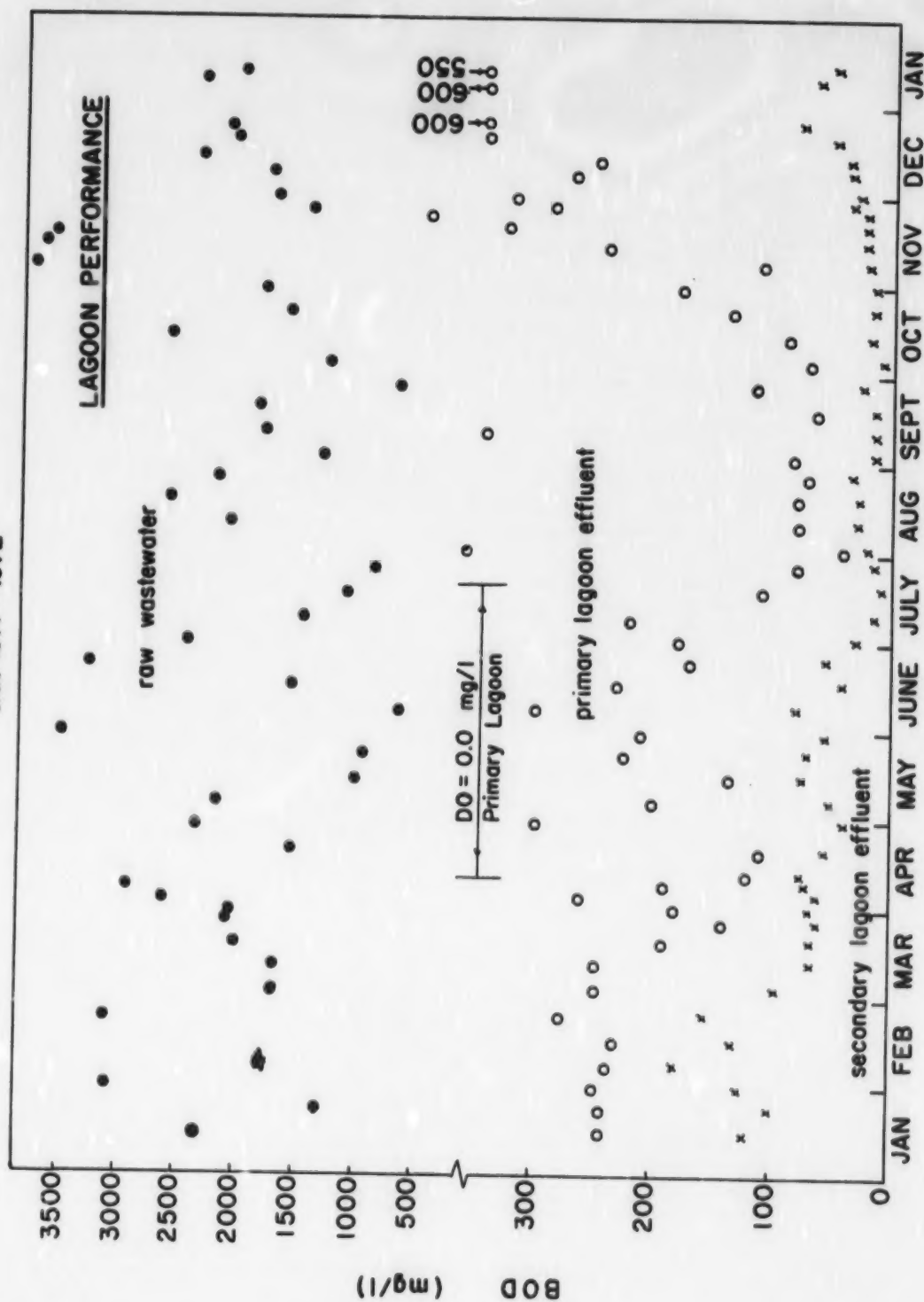


FIGURE 11

TABLE 6
AVERAGE BOD ANALYSES

Quarter 1971	Influent		Primary Effluent		Total	Secondary Effluent	
	Total	Soluble	Total	% R.* (Total)		Soluble	% Overall R.* (Total)
Jan-Mar	1940	1060	224	88.5	106	51	94.5
Apr-June	2040	1110	204	90.0	61	32	97.0
July-Sept.	1530	870	122	92.0	21	13	98.5
Oct-Dec.	2100	1420	274	87.0	31	14	98.5
AVG.	1910	1140	209	89.0	52	27	97.3

All values of BOD in mg/l

*BOD Removal based on total BOD

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

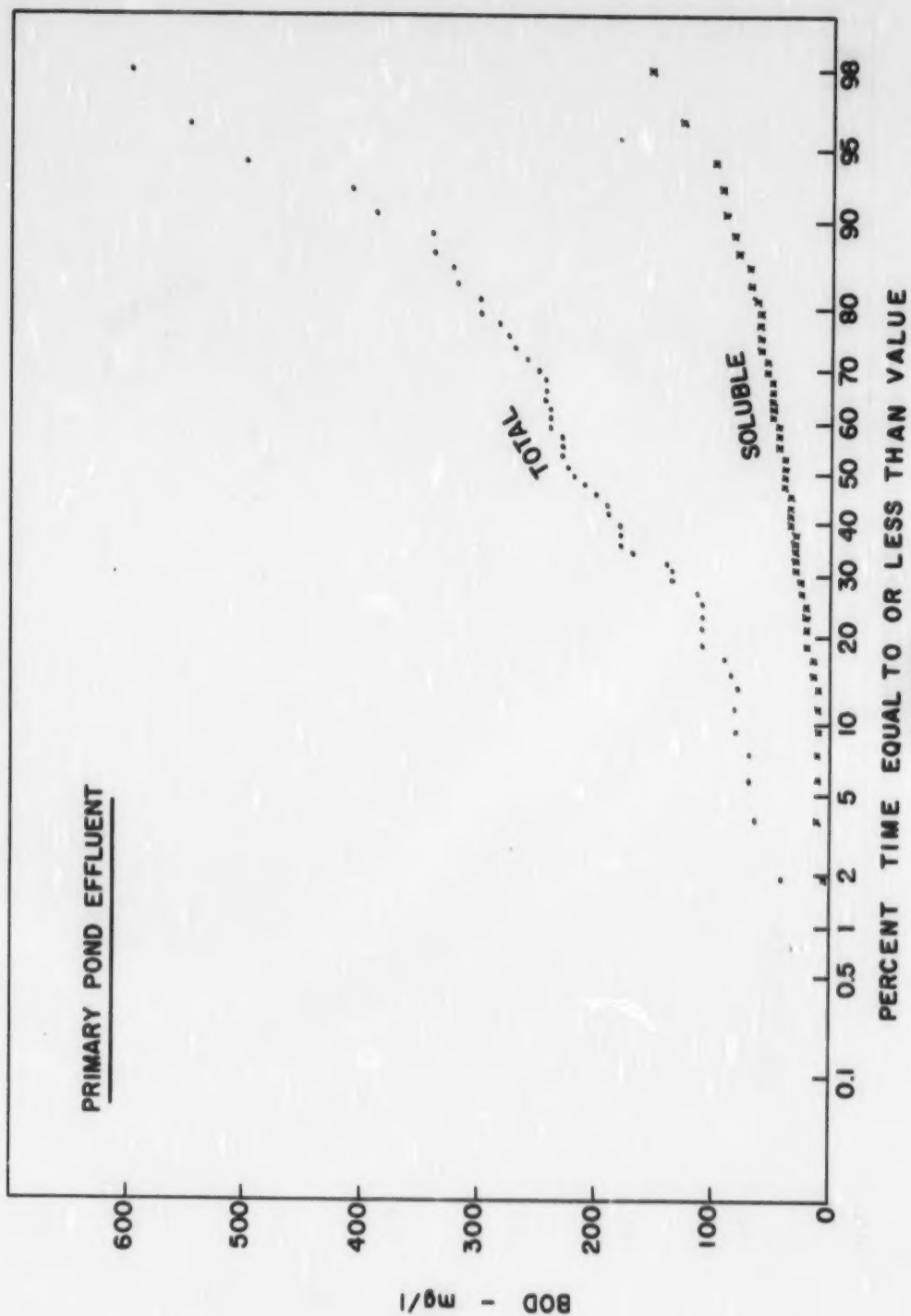


FIGURE 12

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

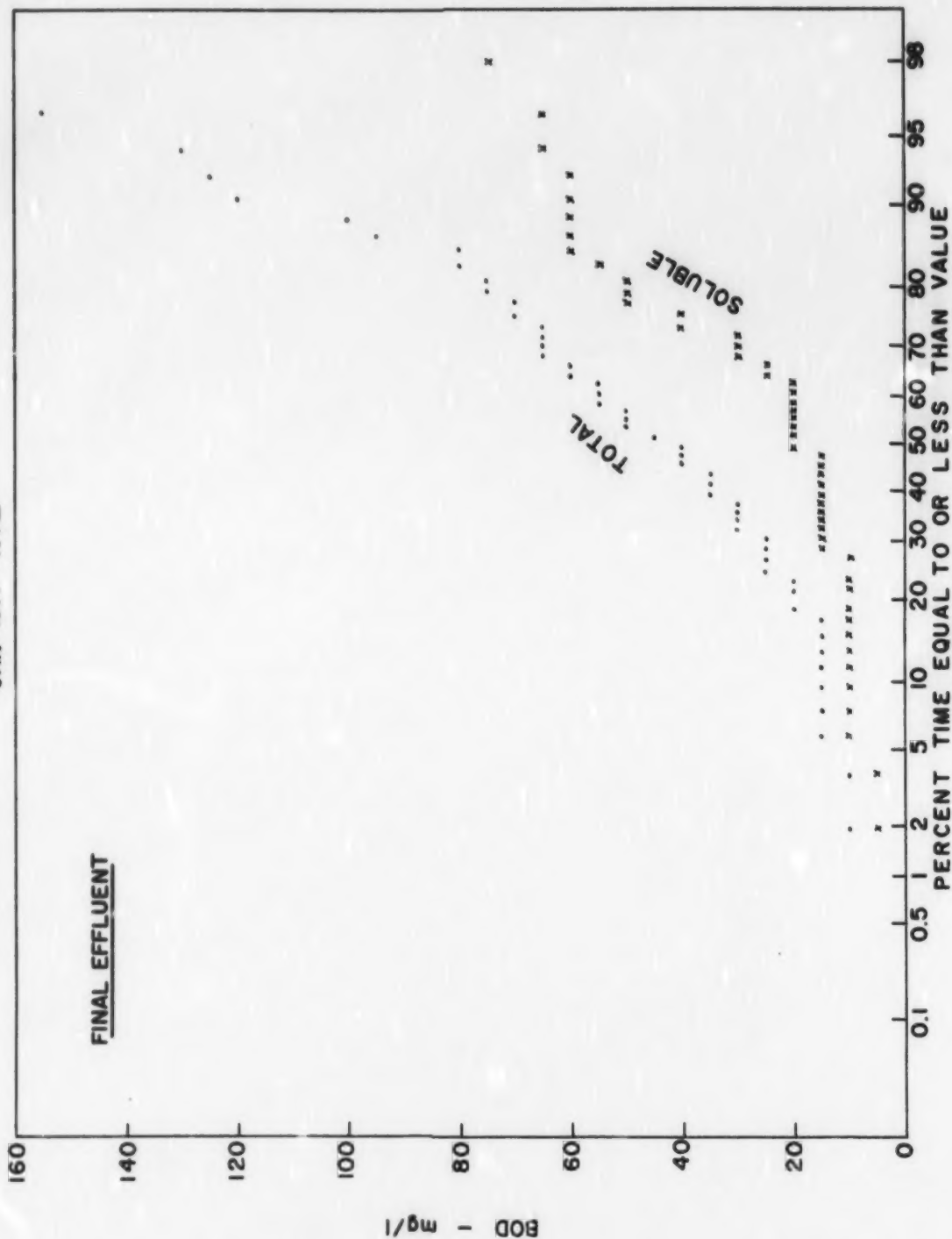


FIGURE 13

system may be inappropriate unless the significant independent variables have been described.

The design parameters most often used for aerated lagoon design include BOD loading per unit volume, detention time, and BOD removal rate (usually expressed as a zero or first order kinetic parameter). For reasons stated above, a simplistic analysis of design parameters is provided. Table 7 summarizes by quarter the three design parameters most often associated with aerated lagoons.

Examination of Tables 6 and 7 suggests that BOD loadings, although increasing slightly during the summer months, are relatively uniform based on quarterly analysis. Day to day variations are of little importance in overall lagoon performance owing to the long detention periods. Note the significant increase in flow during the third quarter (July-Sept) resulting in a reduction in detention time of approximately 15 percent from the average.

The estimation of the first-order removal rate constant is of academic interest only. The assumption of completely-stirred conditions is probably reasonable for the soluble BOD fraction based on oxygen dispersion measurements, but total BOD undoubtedly is greatly influenced by the mixing patterns in the lagoon. No measurements were made to ascertain quantitatively the mixing properties of the lagoon system. A first-order reaction is also assumed, although there was no measurement made in this study to confirm it. First-order kinetics are commonly applied to lagoon systems (8,9,16). The magnitude of the removal rate constants are reasonable for this type of system. The seasonal influence on these rates are difficult to separate from the effects of increased flow rates. In any event, a temperature coefficient θ_t , defined by the relationship

$$k_{T_1} = k_{T_2} \theta_t^{T_1 - T_2} \quad (2)$$

was estimated for the primary lagoon employing the quarterly data presented in Table 7. By solving equation (2) for θ_t , an average value for θ_t for soluble BOD in the primary lagoon was 1.015 whereas a similar value^t for the total BOD was 1.012. These values are substantially lower than those normally reported for aerated lagoon systems (2, 16). Efforts to calculate θ_t for the secondary lagoon were unsuccessful owing to the variation^t of the first-order rates calculated. It is important to note that the influence of temperature on biological processes is related primarily to changes in biochemical reaction rates. This can normally be detected through analysis of oxygen uptake rates, dehydrogenase activity, and other biochemical analyses. As stated earlier, volatile solids measurements in this study were meaningless indicators of biological activity owing to the high volatile solids of an inert nature occurring in the raw wastewater.

TABLE 7

LAGOON LOADING PARAMETERS SUMMARY*

Quarter	Flow (gal/day)	BOD Loading (Total/Soluble)			Volumetric Load		Hydraulic Detention Time		First-Order** BOD Removal Rate/Sec.		Temp.	
		Inf.	Prim. Eff.	Sec. Eff.	Prim.	Sec.	Prim.	Sec.	Prim.	Sec.	Prim.	Sec.
1971	(gal/day)	(lb/day)			(lb/d/1000ft ³)		(Days)		(Days ⁻¹)		(°C)	
Jan-Mar	14,900	270 143	31.2 6.7	14.2 7.2	2.12 1.12	0.24 0.05	64	64	0.12 0.30	0.019 —	2.2	1.0
Apr-June	17,300	285 159	35.9 7.2	10.7 5.2	2.24 1.25	0.28 0.06	55	55	0.13 0.39	0.042 0.007	17.8+	17.8+
July-Sept	20,000	300 169	26.9 8.4	4.6 2.8	2.35 1.33	0.21 0.07	48	48	0.21 0.40	0.100 0.041	21.7	21.7
Oct-Dec	15,200	284 191	35.7 8.5	4.2 8.3	2.22 1.50	0.28 0.07	63	63	0.11 0.34	0.119 0.0003	9.5	9.5
Ave.	16,900	285 166	32.4 7.7	8.4 5.9	2.24 1.30	0.25 0.06	57	57	0.14 0.36	0.05 0.005	—	—

*all values averaged from Appendix C

**k = $\frac{L_0 - L_e}{L_0 \theta}$, based on BOD assuming first-order kinetics, CSTR.

+Estimated, based on estimated temperature for May of 20°C

Suspended Solids - The performance of the lagoon system with respect to suspended solids is considerably more difficult to quantify than BOD. Influent suspended solids were high (Table 8) and contained a large fraction of organic matter. Adequate data was not available on the settling properties of the influent solids and it is apparent from Figure 14 that suspended solids removals in the primary lagoon were very sporadic. During the oxygen deficient period, primary effluent solids often exceeded the influent solids level.

Again the advantage of staged lagoon operation is evident in Figure 14 where it is noted that final effluent suspended solids were considerably more attenuated. The probability plots (Figures 15 and 16) indicate that the effluent suspended solids are not distributed normally, being rather widely scattered at the high end of the distribution. The lowest values of final effluent suspended solids occurred during the period July 1 through September 19. In this period, suspended solids concentrations were less than 50 mg/l, 60 percent of the time. The average suspended solids concentration in the final effluent was 108 mg/l of which 91% was volatile matter (Table 8). It is interesting to note that the solids discharged were predominately volatile, especially during the July through December period. Furthermore, lowest effluent solids occurred during the period when greatest algal growth was evident.

The effluent discharge structure, as noted previously, was a vertical standpipe with no baffling. This discharge pipe was used to control lagoon elevation. Although no data was available, it is reasonable to assume that baffles may more significantly reduce suspended solids discharges.

Solids Contributions to Effluent BOD - The data presented above suggests that suspended solids play a significant role in the bio-degradable organic matter discharged from the lagoons. A simple mathematical expression for this contribution is:

$$L_e(\text{total}) = L_e(\text{soluble}) + C_v S_e \quad (3)$$

where S_e is the concentration of volatile suspended solids in mg/l, and L_e is the total or soluble BOD in mg/l and C_v is the fraction of the volatile solids contributing to the BOD (total).

Plots of effluent insoluble BOD versus volatile suspended solids (VSS) for both primary and secondary lagoons are presented in Figures 17 and 18. Although not well correlated, there is an indication that approximately 46 percent of the VSS contribute to the total BOD in the primary lagoon. This value is high indicating that the primary lagoon solids are not well stabilized. The contribution of VSS to the final effluent BOD is also quite variable, the average value of C_v being 0.29.

TABLE 8
AVERAGE SUSPENDED SOLIDS ANALYSES

Quarter	Influent		Primary Effluent		Secondary Effluent	
	Total	Volatile	Total	Volatile % *	Total	Volatile % Overall*
				(Total)		(Total)
Jan-Mar	658	614	403	335	155	143
Apr-June	600	569	477	390	119	103
July-Sept	547	520	239	197	43	43
Oct-Dec	595	565	445	377	111	110
Average	602	567	395	328	108	98
				38.6		73.4
				20.5		80.1
				56.3		92.1
				25.2		81.3
				34.4		82.0

All values of suspended solids in mg/l

*Percent removal based on total suspended solids

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

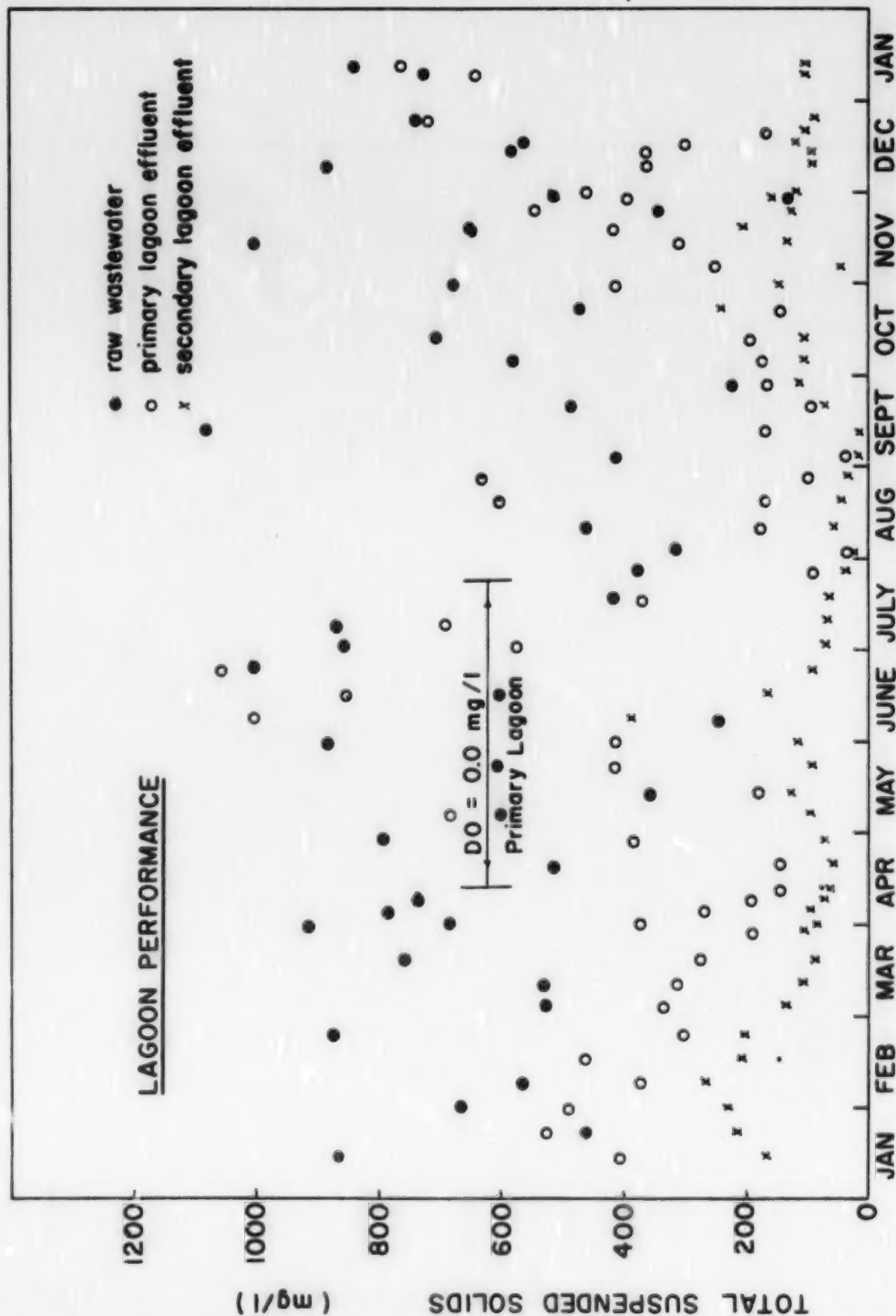


FIGURE 14

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

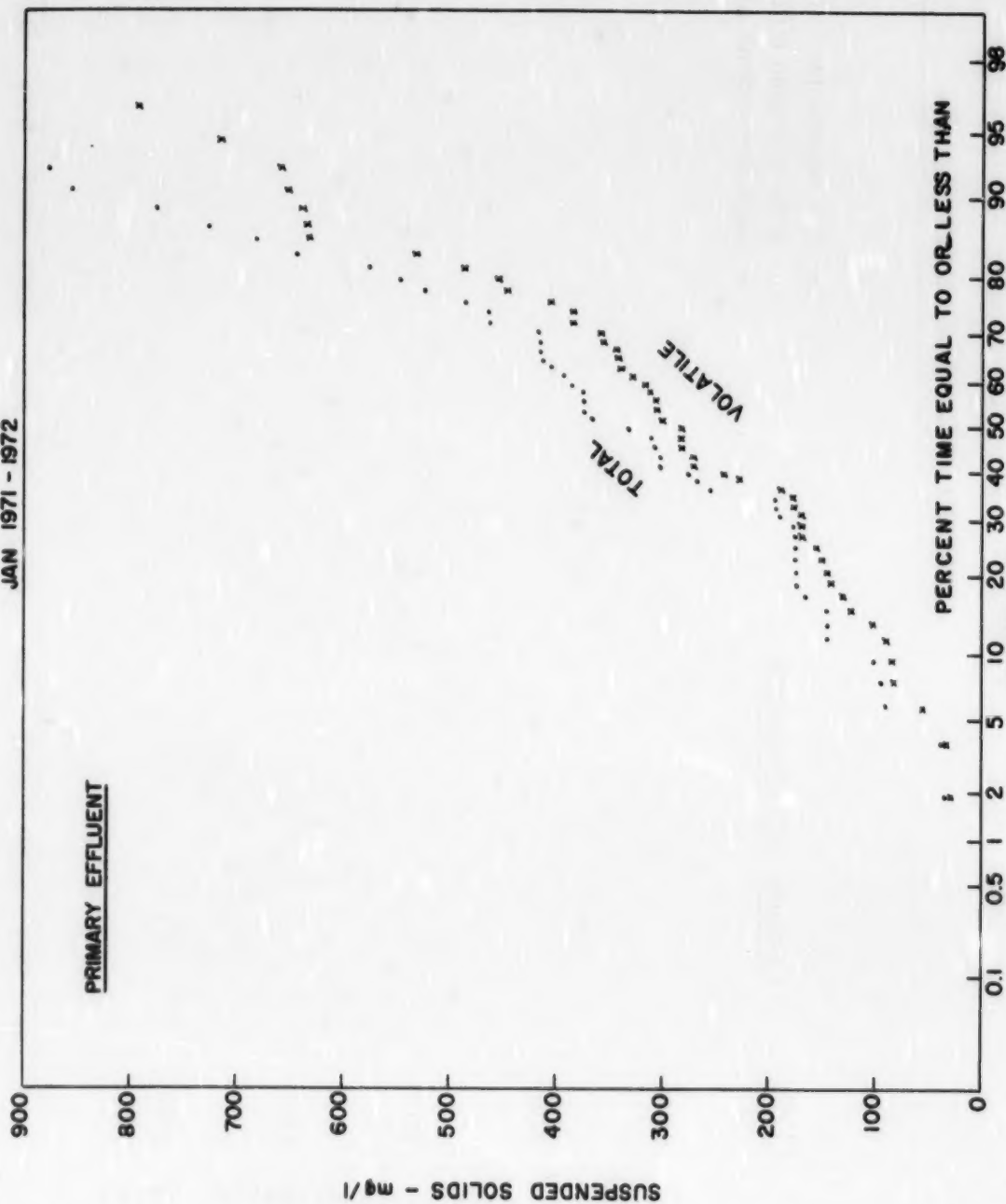


FIGURE 15

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

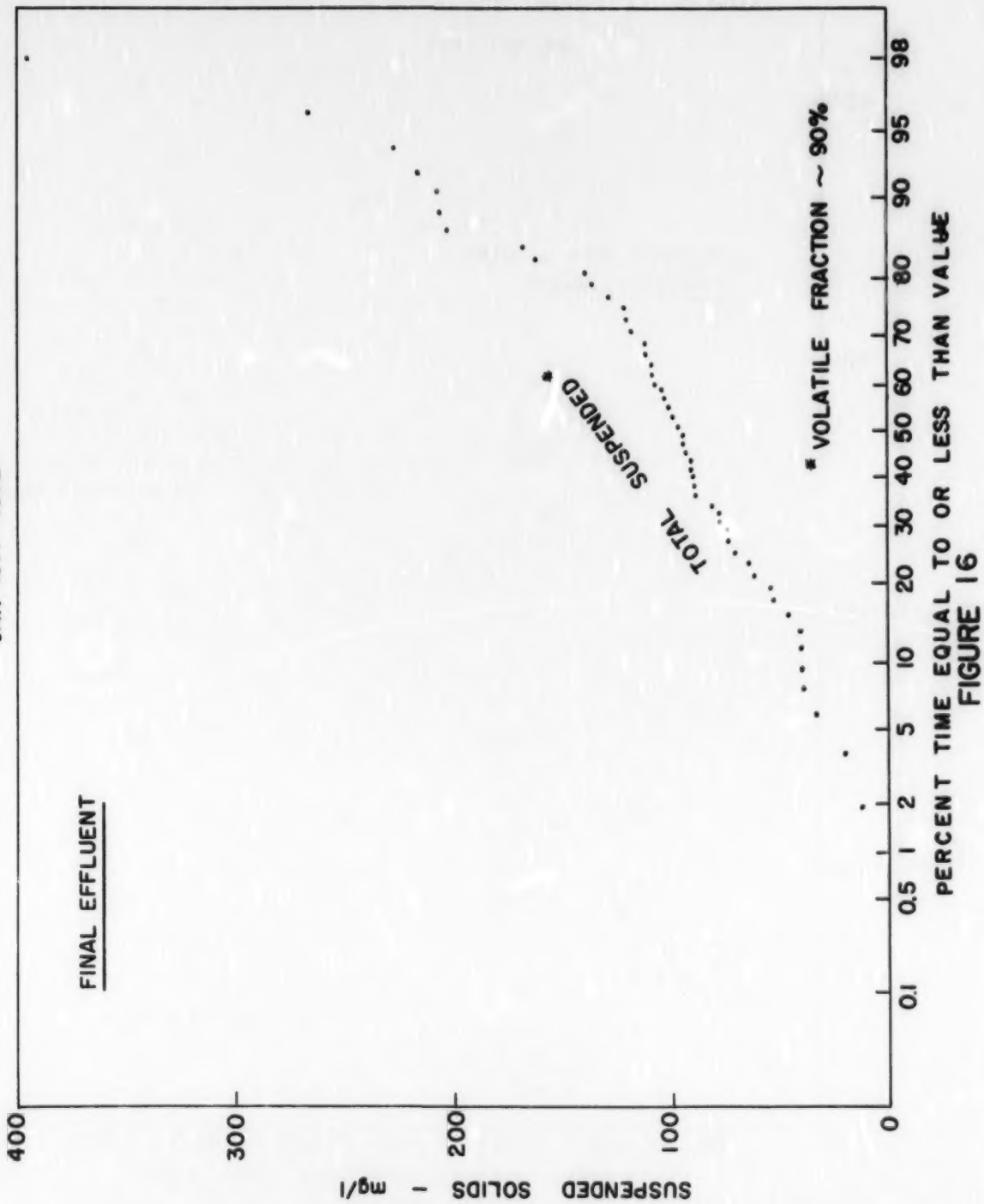


FIGURE 16

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971-1972

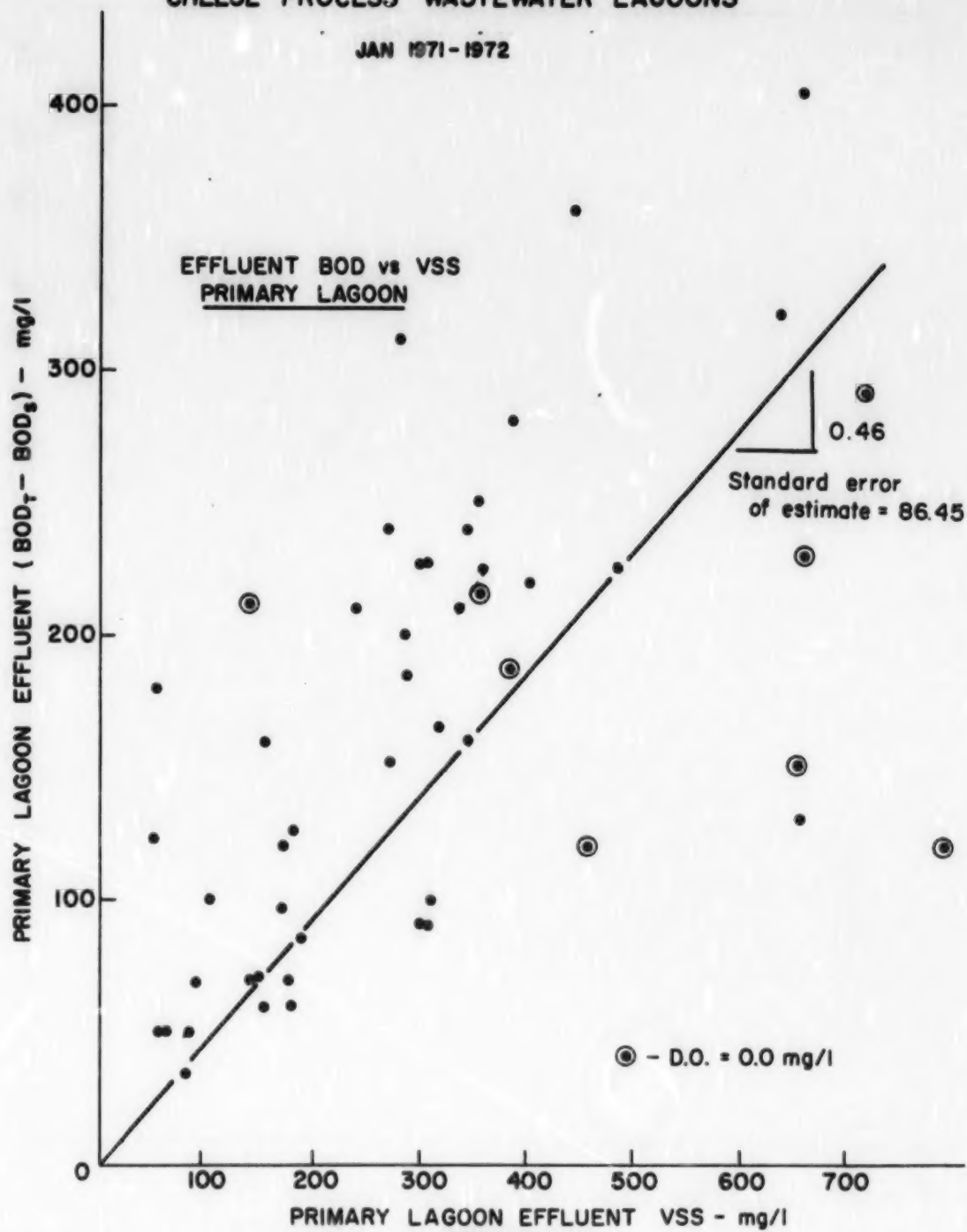


FIGURE 17

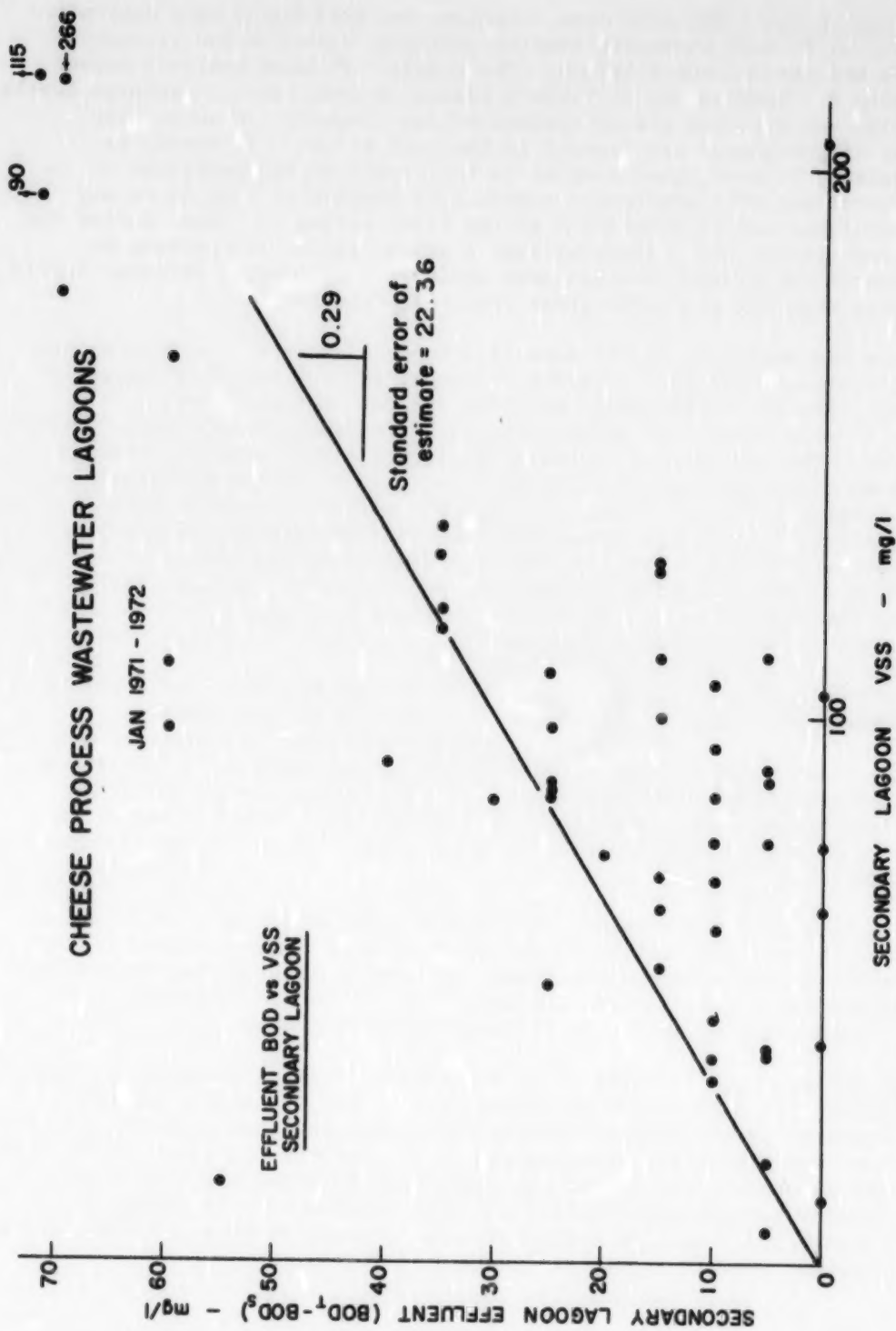


FIGURE 18

Nutrient Levels - The nutrients, nitrogen and phosphorus were determined monthly on 24 hour composite samples collected from the raw wastewater stream and the secondary lagoon. The results of these analyses appear in Table 9. Examination of Table 8 indicates that total phosphorus levels are high owing to the use of phosphate base cleaners. A substantial amount of phosphorus was removed in the pond system. It should be emphasized, however, that removal is the result of sedimentation of the phosphorus and that eventually substantial amounts of phosphorus may be resolubilized and be discharged in the final effluent. Thus, during the one year period, it is doubtful that a steady state with respect to phosphorus transformations has been achieved and, likely, effluent levels may rise significantly over those presently reported.

No data was available on the ammonia nitrogen levels in the system, but it is apparent that total Kjeldahl nitrogen is very low with respect to carbon. The BOD to Nitrogen ratio for the raw wastewater averaged 100:0.7, well below that normally required for adequate biological growth (100:5). The substantial removals of total Kjeldahl nitrogen through the pond system is indicative of both cell synthesis and sedimentation. It is expected that nitrogen recycle would play an important role in the biological system in these lagoons. Nitrate nitrogen in the final effluent was normally low, but warmer temperatures did result in significant nitrification between May and September. An examination of lagoon temperature during these months (Figure 10) indicate that when temperatures exceeded 20°C the rates of nitrification did increase.

Coliform and Fecal Coliform - The sanitary wastewaters from the cheese processing plant were discharged to the process sewers, thereby requiring the installation of chlorination equipment for disinfection of the final effluent. Both total and fecal coliforms were determined from 24 hour flow composited samples of the raw wastewater and final effluent preceding disinfection. Results of these analyses appear in Figures 19 and 20. During this study, no chlorination of the effluent was practiced.

Examination of Figures 19 and 20 indicate that better than 99.9 percent removal of the total coliforms and fecal coliforms was achieved in the lagoons. Temperature did not appear to significantly influence coliform disappearance in the lagoons during the survey period, although there was considerable variation in effluent coliform numbers.

Other Measurements - Values of pH and alkalinity were measured routinely throughout the one year study (Appendix C). Raw wastewater pH values were highest during January through May, averaging approximately 6.5, and then decreasing and leveling off to 6.2 for the remainder of the study. Similarly alkalinities of the raw wastewater averaged 330 mg/l during the January through April period and then continuously decreased to approximately 230 mg/l in September. It is possible that process changes or raw water quality accounted for these changes.

TABLE 9
NITROGEN AND PHOSPHORUS

Primary Lagoon Influent				Secondary Lagoon Effluent								
Date	Day	Total Phosphorus	Total Kjeldahl Nitrogen	Nitrate Nitrogen	BOD:N:P**	Total Phosphorus	Total Kjeldahl Nitrogen	Nitrate Nitrogen				
		mg/l	lb/d*	mg/l		mg/l	lb/d*	mg/l	lb/d*			
1-22	F	37.0	8.5	7.5	1.7	0.1	100:0.6:0.3	35.0	8.0	4.0	0.9	0.3
2-23	Tu	35.0	4.1	9.2	1.1	1.0	100:0.3:1.1	27.5	3.3	5.9	0.7	0.5
3-19	F	41.3	4.9	10.8	1.3	0.1	100:0.5:2.1	26.4	3.1	5.5	0.7	0.2
4-12	M	7.4	0.7	10.2	1.0	0.4	100:6.8:4.9	28.0	2.8	6.0	0.6	0.4
5-14	F	32.5	5.4	5.3	0.9	0.4	100:0.5:3.2	26.8	4.5	10.6	1.8	1.9
6-15	Tu	76.9	13.2	16.0	2.7	4.4	100:1.0:5.0	10.7	1.8	4.2	0.7	24.4
7-9	F	15.5	3.1	12.2	2.4	0.1	100:0.8:1.0	14.8	2.9	1.5	0.3	3.4
8-10	Tu	33.0	5.8	3.6	0.6	0.1	100:0.2:1.6	11.6	2.0	1.0	0.2	4.0
9-11	Sa	68.4	14.0	12.1	2.5	0.1	100:0.7:3.8	8.4	1.7	1.1	0.2	1.4
10-13	W	55.7	8.2	2.0	0.3	0.1	100:1.1:2.2	14.8	2.2	0.2	0.0	0.4
11-14	Su	04.6	14.4	26.2	3.9	0.1	100:0.7:2.6	23.1	3.5	1.8	0.3	0.4
12-16	Th	42.6	5.7	7.7	1.0	2.7	100:0.5:2.7	31.6	4.3	4.2	0.6	0.2
Avg.		45.0	7.3	10.2	1.6	--	100:0.7:2.2	21.6	3.3	3.8	0.6	--

*lb/d - Calculated by using flow rates on day parameter was measured

** BOD:N:P - ratio of weights (or concentrations) of each component on day of analysis

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

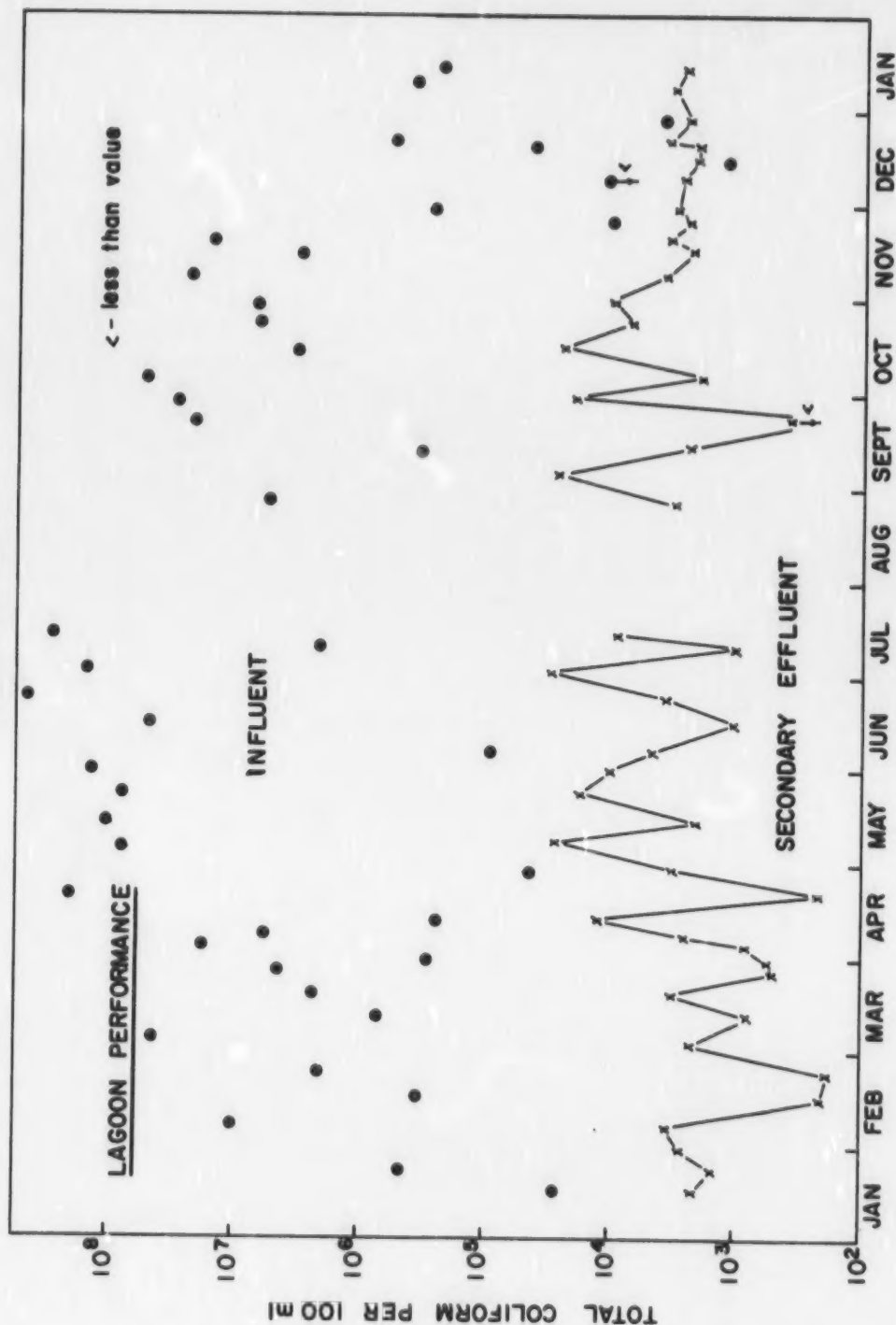


FIGURE 19

CHEESE PROCESS WASTEWATER LAGOONS

JAN 1971 - 1972

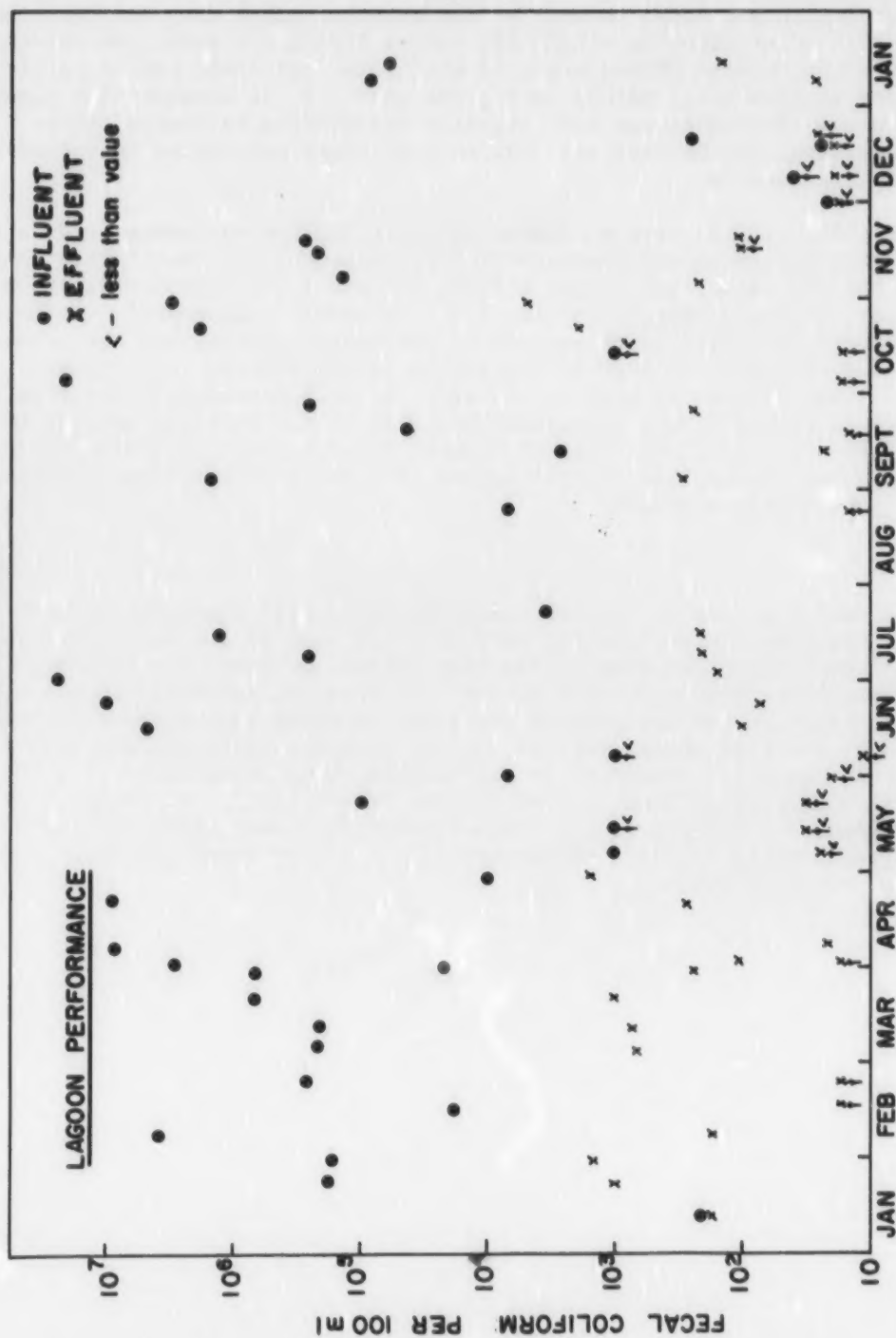


FIGURE 20

The lagoons exhibited pH values of approximately 7.4 during the first one half of the year (January through June), which then sharply decreased to a value of about 7.2 for the remainder of the study. Alkalinities ranged from about 310 mg/l to as high as 540 mg/l during the year. No correlation was noted between pH and measured alkalinity and there was no particular trend in alkalinity values during the year. No correlation was noted between pH and measured alkalinity and there was no particular trend in alkalinity values during the year. It is assumed that changes in algal activities and microorganism respiration influenced these variations, but no data was available on algal activities to document these variations.

Settleable solids were monitored daily on the raw wastewater and on the primary and secondary lagoon effluents (Appendix C). Settleable solids in the raw wastewater ranged between 0.1 and 5 ml/liter during the one year study, normally being less than 1.0. Secondary lagoon effluents never demonstrated significant amounts of settleable solids, but the primary lagoon effluent was high in settleable solids from mid-May through mid-July. Values as high as 120 ml/liter were recorded during this period when dissolved oxygen concentrations were zero. Floating mats of solids and rising sludge was noted throughout this period. Once dissolved oxygen was maintained in this lagoon, the settleable solids decreased to an immeasurable amount.

Costs

The estimated annual capital costs for wastewater treatment at Kent Cheese Co. are based on an interest rate of 7 1/2 percent for mechanical and equipment amortized over a five year period and costs for lagoon construction over a 30 year period. Whereas the estimated annual costs for operation and maintenance are based on costs associated with operating labor, electric power, repairs and replacements, supplies and administration. The annual costs presented do not include those costs associated with the post construction studies that are not representative of normal operating expenses. A summary of the costs related to annual operation for the period January 18, 1971, to January 17, 1972, of these treatment facilities are presented below.

Construction Costs

		Annual Costs
Contract	29,097.42	
Engineering		
Preconstruction Report	500.00	
Preparation of Plans	3,387.14	
Construction Supervision	916.00	
sub total	<u>\$33,900.56</u>	
(i = 7.5% 30 yr)		\$ 2,870.40
Mechanical Equipment	15,592.82	
(i = 7.5% 5 yr)		<u>3,853.99</u>
sub total		<u>\$ 6,724.39</u>
Operating and Maintenance Costs		
Operator labor	\$ 1,364.83	
Electric Power (2.676¢/kwh		
x 92575 kwhr)	2,477.71	
Repairs and Replacement	1,725.43	
Supplies (expendable)	347.47	
Administration	<u>144.62</u>	
sub total	<u>6,060.06</u>	\$ 6,060.06
Total Annual Costs		\$12,784.45

The estimated costs associated with the treatment of wastewaters for the period January 18, 1971, through January 17, 1972, are based on the interest rates and amortization periods selected for fixed costs, whereas the operating and maintenance costs are those costs actually incurred for the period of operation indicated. One item in the Repairs and Replacement costs category; that is the replacement of a single blower with associated costs, accounted for all but \$134.55 of the \$1,725.43 indicated in this category. Normally one would expect a longer period of performance for this type of equipment.

The costs prorated to pounds of BOD applied or 1000 gallons of wastewater treated are \$0.134 per pound of BOD and \$2.05 per 1000 gallons respectively. Based on cheese produced, the estimated cost is \$0.0032 per pound.

SECTION VII
ACKNOWLEDGMENTS

This study was supported by the United States Environmental Protection Agency, Office of Research and Monitoring, and Kent Cheese Company, Inc., Kent, Illinois.

William C. Boyle, Ph.D., and Lawrence B. Polkowski, Ph.D., were contracted as research consultants and authors after the treatment plant was in operation and thereafter were responsible for the interpretation of data collected during the post-construction studies and for writing this report.

Lorne Gramms, Ph.D., Tom Jensen, E.I.T., and Jack Quigley, P.E., were employed as part-time assistants to the project.

Mr. Allen Fehr, P.E., Freeport, Illinois, designed and supervised construction of the treatment plant.

E.P.A. Representatives included William J. Lacy, H. G. Keeler, Robert Burm, and Max Cochrane.

Laboratory analysis were performed by Robert Corning, Chief Chemist, and Mrs. Susan Kelly, Laboratory Supervisor, Corning Laboratories, Cedar Falls, Iowa.

The Assistant Operator, plant operator, was Michael Green, Kent Cheese Company, Kent, Illinois.

Mrs. Kathy Watson and Mrs. Barbara Daul, Lena, Illinois, served as administrative assistants.

SECTION VIII

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SECTION IX
GLOSSARY OF TERMS

Alpha, α	Ratio of overall oxygen mass transfer coefficients in waste to tap water
Beta, β	Ratio of oxygen saturation value in waste to tap water
BOD	Biochemical Oxygen Demand, Five-day, 20°C
C	Oxygen concentration in water, mg/l
C_s	Oxygen saturation concentration in water, mg/l
C_v	Fraction of VSS contributing to BOD
CIP	Clean-in-place
cu ft	Cubic foot
d	Day
DO	Dissolved oxygen concentration
ft	Feet
fps	Feet per second
gal	Gallon
gpd	Gallons per day
hp	Horsepower
k	First order removal rate of BOD, 1/days
k_1	BOD deoxygenation constant
kw	Kilowatt
kw-hr	Kilowatt-hour
L_o	Five-day 20°C BOD of influent, mg/l
L_e	Five-day 20°C BOD of effluent, mg/l

lb	pound
MG	Million gallons
mg/l	Milligrams per liter
N	Nitrogen
N_f	Field oxygen transfer efficiency, $\frac{\text{lb O}_2}{\text{kw-hr}}$
N_s	Standard oxygen transfer efficiency at 20°C, zero D.O. in tap water, $\frac{\text{lb O}_2}{\text{kw-hr}}$
O/R	Over-the-road trucking
O_t	Percent concentration of oxygen in air leaving lagoon, %
P	Phosphorus
P_b	Absolute pressure at point of air release, lb/in^2
Q	Flow rate, MGD
R	Percent removal of polluttional constituent
S_e	VSS Concentration
SS	Suspended solids
Sec	Second
Scfm	Standard cubic foot per minute
θ	Detention time, V/Q , days
θ_t	Temperature correction coefficient
T	Temperature, °C
v	Volts
V	Volume
VSS	Volatile suspended solids

SECTION X

APPENDICES

APPENDIX A
OXYGEN UPTAKE RATES
Primary Lagoon

<u>Date</u>	<u>Station</u> (Figure 4)	<u>Depth</u> from water surface (ft)	<u>Temp</u> °C	<u>Uptake Rate</u> (mg/l/hr)
11-19-70	1	5		
	3	5	6.3	1.9
	5	5	6.3	2.3
	13	5	6.2	2.1
	14	5	6.2	2.0
		5	6.3	1.9
5-23-71	1	1	17.1	5.0
	1	5	17.1	5.8
	5	1	17.1	6.3
	5	6	17.0	6.0
	14	2	17.1	4.9
7-21-71	1	6	24.0	1.2
	5	6	24.1	1.2
	8	6	24.0	1.3
	13	6	24.0	1.1
8-26-71	3	6	23.5	0.9
	5	6	--	1.0
	12	6	23.5	0.8
	14	6	23.5	0.6
10-7-71	1	2	16.5	1.6
	14	6	16.5	1.4
	8	6	16.5	1.6

APPENDIX B

Calculation of Oxygen Transfer Rate

November 19, 1970

Lagoon No. 1 (Primary Lagoon)

Average Oxygen Uptake Rate = 2.1 mg/l/hr = 16.7 lb/hr

Lagoon Temp. = 6.3°C

Dissolved Oxygen = 2.9 mg/l

1 compressor - 30 amps at 220 v = 6.6 kw

Air Distribution to Primary Lagoon - 80% estimated

Field Transfer: (17, 18)

$$N_f = \frac{16.7 \text{ lb/hr}}{6.6 \text{ kw} \times 0.8} = 3.16 \text{ lb/kw-hr}$$

Standard Transfer:

$$N_s = N_f \frac{C_s}{(C_s - C)} \theta^{T-20}$$

$$\begin{aligned} C_s \text{ at } 20^\circ\text{C} &= 9.2 \left(\frac{P_b}{29.4} + \frac{O_t}{42} \right) \\ &= 9.2 \left(\frac{14.7 + 12/2.3}{29.4} + \frac{18.3}{42} \right)^* \end{aligned}$$

$$\begin{aligned} C_{s_{20^\circ}} &= 9.2 (0.68 + 0.5) \\ &= 10.8 \text{ mg/l} \end{aligned}$$

$$C_{s_{6.3^\circ\text{C}}} = 14.8 \text{ mg/l @ 12' depth}$$

$$\alpha = 0.75, \beta = 0.96, \theta = 1.02^{(19)}$$

$$N_s = 3.16 \left[\frac{10.8}{0.75(0.96 \times 14.8 - 2.8) 1.02^{6.3-20}} \right]$$

$$N_s = 5.25 \text{ lb/kw-hr}$$

$$N_s = 3.91 \text{ lb/hp-hr}$$

*where depth = 12' and estimated transfer efficiency is estimated at 10%

APPENDIX C

Lagoon Data

WEEKLY DATA

DATE	DAY	LAGOON INFLUENT (GPD)	BOD ₅ (TOTAL/SOLUBLE)						TOTAL SUSPENDED SOLIDS						VOLATILE SUSPENDED SOLIDS						TOTAL COLIFORMS/SPECIAL COLIFORMS					
			S.P.1	S.P.2	S.P.3	S.P.1	S.P.2	S.P.3	S.P.1	S.P.2	S.P.3	S.P.1	S.P.2	S.P.3	S.P.1	S.P.2	S.P.3	S.P.1	S.P.2	S.P.3	S.P.1	S.P.2	S.P.3	S.P.1	S.P.2	S.P.3
1-14	TH	21,000	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
1-22	F	21,500	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
1-30	S	11,800	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
2-7	SUN	10,300	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
2-15	M	13,200	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
2-23	T	14,200	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
3-3	W	17,500	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
3-11	TH	13,300	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
3-19	F	14,300	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
3-27	S	18,000	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
3-31	W	25,000	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
4-4	SUN	16,700	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
4-8	TH	19,300	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
4-12	M	32,500	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
4-20	W	16,200	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
5-6	TH	17,500	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
5-14	F	20,000	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
5-22	S	11,600	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
5-30	SUN	—	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
6-7	M	43,100	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
6-15	T	20,600	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
6-23	W	18,400	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
7-1	TH	25,900	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
7-9	F	23,600	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
7-17	S	21,700	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
7-25	SUN	—	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
8-2	M	30,000	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
8-10	T	21,100	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
8-18	W	19,700	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
8-26	TH	21,000	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
9-3	F	21,800	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
9-11	S	24,600	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
9-19	SUN	—	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
9-27	M	33,900	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
10-5	T	23,400	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100
10-13	W	19,700	1340	410	110	84	151	405	71	167	29.2	846	148	340	575	167	29.2	280	280	280	100	100	100	100	100	100

WEEKLY DATA

DATE	DAY	LAGOON INFLUENT (GPD)	BODs			(TOTAL SOLUBLE)			TOTAL SUSPENDED SOLIDS			VOLATILE SUSPENDED SOLIDS			TOTAL COLIFORMS/100 ML		
			S.R. 1	S.R. 2	S.R. 3	S.R. 1	S.R. 2	S.R. 3	S.R. 1	S.R. 2	S.R. 3	S.R. 1	S.R. 2	S.R. 3	S.R. 1	S.R. 2	S.R. 3
10-21	TH	23,200	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
10-29	F	17,700	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
11-6	S	15,400	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
11-14	SUN	18,200	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
11-18	TH	15,015	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
11-24	W	15,700	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
11-26	F	14,200	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
11-30	T	17,000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
12-8	W	19,310	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
12-13	M	9,675	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
12-16	TH	16,155	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
12-20	M	10,325	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
12-24	F	17,605	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
1-9	SUN	11,130	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000
1-12	W	17,415	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000	1000

To: C. W. Klassen, Director
Environmental Protection Agency
Bureau of Water Pollution Control
825 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

[illegible]

To: C. W. Klamm, Director
Environmental Protection Agency
Division of Water Pollution Control
1215 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

Comments	DATE	PLANT PROCESS		Plant Flow (MGD)	Plant Flow (MGD)	Plant Flow (MGD)	SETTLABLE SOLIDS			ALUMINUM			WEATHER
		IN	OUT				1	2	3	1	2	3	
FEBRUARY 1971 DAILY LOG 2:11 AM OFFSHORE OF 2nd CELL	1 M	13,062	28,000	2100	2100	1	—	—	6.7	7.4	7.5	484	70.32 80.32
	2 T	11,883	17,500	2100	15,910	1	—	—	6.6	7.4	7.4	410	71.32 80.32
	3 W	12,443	17,000	2100	16,200	3	—	—	6.7	7.3	7.4	470	70.32 80.32
	4 T	9,730	17,100	2100	15,800	2	—	—	6.4	7.4	7.4	440	71.32 80.32
	5 F	10,502	15,900	1800	14,900	1	—	—	6.7	7.4	7.5	480	70.32 80.32
	6 S	8,381	14,600	2100	13,300	1	—	—	6.4	7.4	7.5	484	70.32 80.32
	7 S	15,078	17,500	2100	16,200	2	—	—	7.4	7.4	7.6	502	71.32 80.32
	8 M	13,355	22,100	2100	18,500	2	—	—	6.7	7.5	7.4	484	70.32 80.32
	9 T	13,380	14,400	2100	15,300	2	—	—	6.5	7.4	7.6	444	71.32 79.32
	10 W	9,438	14,900	2100	13,600	4.5	—	—	5.5	7.4	7.4	450	71.32 79.32
GENERAL TYPICAL	11 T	13,831	17,500	1800	16,500	3	—	—	6.2	7.4	7.4	496	71.32 79.32
	12 F	11,606	12,700	2100	11,400	7	—	—	6.4	7.5	7.5	494	71.32 79.32
	13 S	15,614	14,500	2100	13,200	—	—	—	6.3	7.4	7.5	440	71.32 79.32
	14 M	15,974	14,600	2100	15,000	2	—	—	6.6	7.5	7.5	476	71.32 79.32
	15 T	11,222	21,500	2100	20,200	1	—	—	6.5	7.4	7.5	480	71.32 79.32
	16 T	11,035	14,300	2100	12,000	1	—	—	6.5	7.4	7.4	500	71.32 79.32
	17 F	11,603	17,600	1800	14,600	—	—	—	6.3	7.4	7.4	486	71.32 79.32
	18 S	9,475	13,100	2100	11,800	—	—	—	6.4	7.5	7.5	436	71.32 79.32
	19 S	11,478	22,800	2100	21,500	—	—	—	6.3	7.4	7.4	414	71.32 79.32
	20 M	15,878	15,800	2100	14,200	1.5	—	—	6.4	7.4	7.5	416	71.32 79.32
GENERAL TYPICAL	21 W	10,454	14,000	2100	12,700	1	—	—	6.5	7.4	7.5	386	71.32 79.32
	22 T	8,119	17,500	2100	16,200	1	—	—	6.6	7.4	7.5	384	71.32 79.32
	23 F	12,632	13,300	1800	12,100	1	—	—	6.4	7.4	7.4	386	71.32 79.32
	24 S	9,107	14,500	2100	13,200	5	—	—	—	—	—	380	71.32 79.32
	25 S	—	—	—	—	—	—	—	—	—	—	—	71.32 79.32
	26 M	285,340	410,700	5040	379,500	30.4	—	—	41.5	171	111.7	1152	71.32 79.32
	27 T	11,889	14,400	1800	13,350	1.5	—	—	6.2	7.1	7.2	416	71.32 79.32
	28 W	15,974	26,000	2100	24,700	5	—	—	7.4	7.6	7.6	500	71.32 79.32
	29 T	8,119	12,700	1800	11,400	2	—	—	5.4	7.3	7.3	362	71.32 79.32
	30 F	—	—	—	—	—	—	—	—	—	—	—	71.32 79.32
SUMMARY		285,340				30.4				41.5		1152	
TOTAL		285,340				30.4				41.5		1152	
AVERAGE		11,814				1.21				1.71		416	
MAXIMUM		285,340				30.4				41.5		1152	
MINIMUM		8,119				1				6.2		362	
REMARKS		REMARKS FROM 285,340 MONDAY'S											

To: C. W. Mann, Director
Environmental Protection Agency
Bureau of Water Pollution Control
125 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

Comments	DATE	PLANT PROCESS		Flow (GPM)	ROILER LACTO- FLAV (GPM)	SETTLABLE SOLIDS			ALKALINITY			ABSORBANCE			WEATHER
		Run Pkts.	Chem. Pkts.			1	2	3	1	2	3	1	2	3	
MARCH 1971	1 M	15,193	21,500	21,500	21,500	—	—	—	6.6	7.5	7.5	300	370	386	10
	2 T	17,481	18,200	18,200	18,200	1.6	—	—	6.6	7.5	7.5	300	356	374	20
	3 M	12,038	18,200	18,200	18,200	1.5	—	—	6.5	7.4	7.5	300	364	380	14
	4 T	10,513	15,300	15,300	15,300	1	—	—	6.7	7.45	7.4	300	352	380	8
THAWING ICE	5 F	14,731	17,100	18,000	18,000	.2	—	—	6.7	7.5	7.5	300	360	380	0
"	6 S	11,774	15,300	21,000	14,000	2	—	—	6.6	7.4	7.5	296	362	384	30
"	7 S														
THAWING ICE	8 M	14,181	20,400	20,400	19,300	.1	—	—	6.5	7.4	7.5	290	374	390	10
"	9 T	14,804	15,300	24,000	13,700	1	—	—	6.4	7.5	7.5	300	370	386	15
"	10 M	13,929	14,700	21,000	13,400	1	—	—	6.3	7.5	7.5	300	370	388	25
"	11 T	8,747	14,600	21,000	13,300	.6	—	—	6.4	7.45	7.5	296	370	390	30
"	12 F	15,297	14,500	18,000	13,500	1	—	—	6.4	7.5	7.4	294	374	390	30
THAWING ICE	13 S	11,404	15,600	21,000	14,500	2	—	—	6.5	7.4	7.4	300	370	390	34
"	14 S														
THAWING ICE	15 M	16,640	25,800	21,000	24,300	—	—	—	6.4	7.45	7.5	290	374	390	45
"	16 T	17,336	17,100	24,000	15,500	1.5	—	—	6.4	7.4	7.5	298	370	390	24
"	17 M	13,313	17,500	21,000	16,200	1	—	—	6.3	7.45	7.5	305	370	390	18
"	18 T	12,337	14,500	21,000	13,200	1.3	—	—	6.7	7.35	7.4	320	370	380	36
HEAVY SNOW	19 F	11,520	15,300	18,000	14,300	1	—	—	6.6	7.4	7.4	310	374	380	30
"	20 S	12,514	13,900	20,000	12,900	1.2	—	—	6.5	7.4	7.5	315	372	376	30
THAWING COND.	21 S														
BRIGHT YELLOW	22 M	15,243	24,500	21,000	23,200	—	—	—	6.7	7.55	7.5	320	380	376	30
SUM APPEARED	23 T	17,458	24,700	24,000	21,400	1.5	—	—	6.6	7.7	7.5	320	364	390	15
ON SCALES	24 M	15,723	17,100	21,000	15,800	2	—	—	6.6	7.6	7.5	310	364	390	10
"	25 T	20,592	19,300	21,000	18,400	1.6	—	—	6.5	7.5	7.5	340	380	390	20
300 MG P.P.O. CELL	26 F	19,846	18,000	18,000	17,000	1.5	—	—	6.6	7.5	7.5	330	376	370	30
" " " "	27 S	14,396	19,300	21,000	18,000	2.0	—	—	6.5	7.4	7.45	316	376	380	30
ATTORNEY 1" CELL	28 S														
LAST 1/2 OF 204 CELL	29 M	18,309	24,700	21,000	25,400	.5	—	—	6.5	7.5	7.4	330	388	376	30
"	30 T	17,416	20,600	24,000	19,000	2.2	—	—	6.6	7.45	7.55	300	370	374	30
"	31 M	16,354	20,300	21,000	25,000	2.0	—	—	6.7	7.65	7.7	308	388	396	45
"		53,793	50,100	57,000	46,500	2.1	—	—	6.8	7.6	7.6	844	10,080	10,132	48
"		19,923	16,165	18,311	13,019	1.2	—	—	6.5	7.5	7.5	315	371	376	25
"		20,572	26,700	24,000	25,400	2.2	—	—	6.7	7.7	7.7	380	370	390	45
"		8,747	13,600	18,000	12,300	.1	—	—	6.3	7.35	7.4	290	352	346	0
* (WELL-BOILER + BOO (440 INFL))															
* (SUNDAY FLOW DUE IN MONDAY)															

To: C. W. Klamm, Director
Environmental Protection Agency
Bureau of Water Pollution Control
625 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

DATE	TIME	CLOUDS	PLANT PROGRESS		PLANT HEIGHT (cm)	PLANT WEIGHT (g)	PLANT TEMPERATURE (°C)	PLANT HUMIDITY (%)	PLANT COLOR	PLANT STRENGTH	PLANT DISEASE	PLANT INJURY	PLANT DAMAGE	PLANT LOSS (%)	PLANT RECOVERY (%)	PLANT SURVIVAL (%)	PLANT GROWTH (%)	PLANT YIELD (%)	PLANT QUALITY (%)	PLANT VALUE (%)	PLANT COST (%)	PLANT PROFIT (%)	PLANT RISK (%)	PLANT RETURN (%)	PLANT INVESTMENT (%)	PLANT ROI (%)	PLANT NPV (%)	PLANT IRR (%)	PLANT MIRR (%)	PLANT MOIC (%)	PLANT MOC (%)	PLANT MOC2 (%)	PLANT MOC3 (%)	PLANT MOC4 (%)	PLANT MOC5 (%)	PLANT MOC6 (%)	PLANT MOC7 (%)	PLANT MOC8 (%)	PLANT MOC9 (%)	PLANT MOC10 (%)	PLANT MOC11 (%)	PLANT MOC12 (%)	PLANT MOC13 (%)	PLANT MOC14 (%)	PLANT MOC15 (%)	PLANT MOC16 (%)	PLANT MOC17 (%)	PLANT MOC18 (%)	PLANT MOC19 (%)	PLANT MOC20 (%)	PLANT MOC21 (%)	PLANT MOC22 (%)	PLANT MOC23 (%)	PLANT MOC24 (%)	PLANT MOC25 (%)	PLANT MOC26 (%)	PLANT MOC27 (%)	PLANT MOC28 (%)	PLANT MOC29 (%)	PLANT MOC30 (%)	PLANT MOC31 (%)	PLANT MOC32 (%)	PLANT MOC33 (%)	PLANT MOC34 (%)	PLANT MOC35 (%)	PLANT MOC36 (%)	PLANT MOC37 (%)	PLANT MOC38 (%)	PLANT MOC39 (%)	PLANT MOC40 (%)	PLANT MOC41 (%)	PLANT MOC42 (%)	PLANT MOC43 (%)	PLANT MOC44 (%)	PLANT MOC45 (%)	PLANT MOC46 (%)	PLANT MOC47 (%)	PLANT MOC48 (%)	PLANT MOC49 (%)	PLANT MOC50 (%)	PLANT MOC51 (%)	PLANT MOC52 (%)	PLANT MOC53 (%)	PLANT MOC54 (%)	PLANT MOC55 (%)	PLANT MOC56 (%)	PLANT MOC57 (%)	PLANT MOC58 (%)	PLANT MOC59 (%)	PLANT MOC60 (%)	PLANT MOC61 (%)	PLANT MOC62 (%)	PLANT MOC63 (%)	PLANT MOC64 (%)	PLANT MOC65 (%)	PLANT MOC66 (%)	PLANT MOC67 (%)	PLANT MOC68 (%)	PLANT MOC69 (%)	PLANT MOC70 (%)	PLANT MOC71 (%)	PLANT MOC72 (%)	PLANT MOC73 (%)	PLANT MOC74 (%)	PLANT MOC75 (%)	PLANT MOC76 (%)	PLANT MOC77 (%)	PLANT MOC78 (%)	PLANT MOC79 (%)	PLANT MOC80 (%)	PLANT MOC81 (%)	PLANT MOC82 (%)	PLANT MOC83 (%)	PLANT MOC84 (%)	PLANT MOC85 (%)	PLANT MOC86 (%)	PLANT MOC87 (%)	PLANT MOC88 (%)	PLANT MOC89 (%)	PLANT MOC90 (%)	PLANT MOC91 (%)	PLANT MOC92 (%)	PLANT MOC93 (%)	PLANT MOC94 (%)	PLANT MOC95 (%)	PLANT MOC96 (%)	PLANT MOC97 (%)	PLANT MOC98 (%)	PLANT MOC99 (%)	PLANT MOC100 (%)	PLANT MOC101 (%)	PLANT MOC102 (%)	PLANT MOC103 (%)	PLANT MOC104 (%)	PLANT MOC105 (%)	PLANT MOC106 (%)	PLANT MOC107 (%)	PLANT MOC108 (%)	PLANT MOC109 (%)	PLANT MOC110 (%)	PLANT MOC111 (%)	PLANT MOC112 (%)	PLANT MOC113 (%)	PLANT MOC114 (%)	PLANT MOC115 (%)	PLANT MOC116 (%)	PLANT MOC117 (%)	PLANT MOC118 (%)	PLANT MOC119 (%)	PLANT MOC120 (%)	PLANT MOC121 (%)	PLANT MOC122 (%)	PLANT MOC123 (%)	PLANT MOC124 (%)	PLANT MOC125 (%)	PLANT MOC126 (%)	PLANT MOC127 (%)	PLANT MOC128 (%)	PLANT MOC129 (%)	PLANT MOC130 (%)	PLANT MOC131 (%)	PLANT MOC132 (%)	PLANT MOC133 (%)	PLANT MOC134 (%)	PLANT MOC135 (%)	PLANT MOC136 (%)	PLANT MOC137 (%)	PLANT MOC138 (%)	PLANT MOC139 (%)	PLANT MOC140 (%)	PLANT MOC141 (%)	PLANT MOC142 (%)	PLANT MOC143 (%)	PLANT MOC144 (%)	PLANT MOC145 (%)	PLANT MOC146 (%)	PLANT MOC147 (%)	PLANT MOC148 (%)	PLANT MOC149 (%)	PLANT MOC150 (%)	PLANT MOC151 (%)	PLANT MOC152 (%)	PLANT MOC153 (%)	PLANT MOC154 (%)	PLANT MOC155 (%)	PLANT MOC156 (%)	PLANT MOC157 (%)	PLANT MOC158 (%)	PLANT MOC159 (%)	PLANT MOC160 (%)	PLANT MOC161 (%)	PLANT MOC162 (%)	PLANT MOC163 (%)	PLANT MOC164 (%)	PLANT MOC165 (%)	PLANT MOC166 (%)	PLANT MOC167 (%)	PLANT MOC168 (%)	PLANT MOC169 (%)	PLANT MOC170 (%)	PLANT MOC171 (%)	PLANT MOC172 (%)	PLANT MOC173 (%)	PLANT MOC174 (%)	PLANT MOC175 (%)	PLANT MOC176 (%)	PLANT MOC177 (%)	PLANT MOC178 (%)	PLANT MOC179 (%)	PLANT MOC180 (%)	PLANT MOC181 (%)	PLANT MOC182 (%)	PLANT MOC183 (%)	PLANT MOC184 (%)	PLANT MOC185 (%)	PLANT MOC186 (%)	PLANT MOC187 (%)	PLANT MOC188 (%)	PLANT MOC189 (%)	PLANT MOC190 (%)	PLANT MOC191 (%)	PLANT MOC192 (%)	PLANT MOC193 (%)	PLANT MOC194 (%)	PLANT MOC195 (%)	PLANT MOC196 (%)	PLANT MOC197 (%)	PLANT MOC198 (%)	PLANT MOC199 (%)	PLANT MOC200 (%)	PLANT MOC201 (%)	PLANT MOC202 (%)	PLANT MOC203 (%)	PLANT MOC204 (%)	PLANT MOC205 (%)	PLANT MOC206 (%)	PLANT MOC207 (%)	PLANT MOC208 (%)	PLANT MOC209 (%)	PLANT MOC210 (%)	PLANT MOC211 (%)	PLANT MOC212 (%)	PLANT MOC213 (%)	PLANT MOC214 (%)	PLANT MOC215 (%)	PLANT MOC216 (%)	PLANT MOC217 (%)	PLANT MOC218 (%)	PLANT MOC219 (%)	PLANT MOC220 (%)	PLANT MOC221 (%)	PLANT MOC222 (%)	PLANT MOC223 (%)	PLANT MOC224 (%)	PLANT MOC225 (%)	PLANT MOC226 (%)	PLANT MOC227 (%)	PLANT MOC228 (%)	PLANT MOC229 (%)	PLANT MOC230 (%)	PLANT MOC231 (%)	PLANT MOC232 (%)	PLANT MOC233 (%)	PLANT MOC234 (%)	PLANT MOC235 (%)	PLANT MOC236 (%)	PLANT MOC237 (%)	PLANT MOC238 (%)	PLANT MOC239 (%)	PLANT MOC240 (%)	PLANT MOC241 (%)	PLANT MOC242 (%)	PLANT MOC243 (%)	PLANT MOC244 (%)	PLANT MOC245 (%)	PLANT MOC246 (%)	PLANT MOC247 (%)	PLANT MOC248 (%)	PLANT MOC249 (%)	PLANT MOC250 (%)	PLANT MOC251 (%)	PLANT MOC252 (%)	PLANT MOC253 (%)	PLANT MOC254 (%)	PLANT MOC255 (%)	PLANT MOC256 (%)	PLANT MOC257 (%)	PLANT MOC258 (%)	PLANT MOC259 (%)	PLANT MOC260 (%)	PLANT MOC261 (%)	PLANT MOC262 (%)	PLANT MOC263 (%)	PLANT MOC264 (%)	PLANT MOC265 (%)	PLANT MOC266 (%)	PLANT MOC267 (%)	PLANT MOC268 (%)	PLANT MOC269 (%)	PLANT MOC270 (%)	PLANT MOC271 (%)	PLANT MOC272 (%)	PLANT MOC273 (%)	PLANT MOC274 (%)	PLANT MOC275 (%)	PLANT MOC276 (%)	PLANT MOC277 (%)	PLANT MOC278 (%)	PLANT MOC279 (%)	PLANT MOC280 (%)	PLANT MOC281 (%)	PLANT MOC282 (%)	PLANT MOC283 (%)	PLANT MOC284 (%)	PLANT MOC285 (%)	PLANT MOC286 (%)	PLANT MOC287 (%)	PLANT MOC288 (%)	PLANT MOC289 (%)	PLANT MOC290 (%)	PLANT MOC291 (%)	PLANT MOC292 (%)	PLANT MOC293 (%)	PLANT MOC294 (%)	PLANT MOC295 (%)	PLANT MOC296 (%)	PLANT MOC297 (%)	PLANT MOC298 (%)	PLANT MOC299 (%)	PLANT MOC300 (%)	PLANT MOC301 (%)	PLANT MOC302 (%)	PLANT MOC303 (%)	PLANT MOC304 (%)	PLANT MOC305 (%)	PLANT MOC306 (%)	PLANT MOC307 (%)	PLANT MOC308 (%)	PLANT MOC309 (%)	PLANT MOC310 (%)	PLANT MOC311 (%)	PLANT MOC312 (%)	PLANT MOC313 (%)	PLANT MOC314 (%)	PLANT MOC315 (%)	PLANT MOC316 (%)	PLANT MOC317 (%)	PLANT MOC318 (%)	PLANT MOC319 (%)	PLANT MOC320 (%)	PLANT MOC321 (%)	PLANT MOC322 (%)	PLANT MOC323 (%)	PLANT MOC324 (%)	PLANT MOC325 (%)	PLANT MOC326 (%)	PLANT MOC327 (%)	PLANT MOC328 (%)	PLANT MOC329 (%)	PLANT MOC330 (%)	PLANT MOC331 (%)	PLANT MOC332 (%)	PLANT MOC333 (%)	PLANT MOC334 (%)	PLANT MOC335 (%)	PLANT MOC336 (%)	PLANT MOC337 (%)	PLANT MOC338 (%)	PLANT MOC339 (%)	PLANT MOC340 (%)	PLANT MOC341 (%)	PLANT MOC342 (%)	PLANT MOC343 (%)	PLANT MOC344 (%)	PLANT MOC345 (%)	PLANT MOC346 (%)	PLANT MOC347 (%)	PLANT MOC348 (%)	PLANT MOC349 (%)	PLANT MOC350 (%)	PLANT MOC351 (%)	PLANT MOC352 (%)	PLANT MOC353 (%)	PLANT MOC354 (%)	PLANT MOC355 (%)	PLANT MOC356 (%)	PLANT MOC357 (%)	PLANT MOC358 (%)	PLANT MOC359 (%)	PLANT MOC360 (%)	PLANT MOC361 (%)	PLANT MOC362 (%)	PLANT MOC363 (%)	PLANT MOC364 (%)	PLANT MOC365 (%)	PLANT MOC366 (%)	PLANT MOC367 (%)	PLANT MOC368 (%)	PLANT MOC369 (%)	PLANT MOC370 (%)	PLANT MOC371 (%)	PLANT MOC372 (%)	PLANT MOC373 (%)	PLANT MOC374 (%)	PLANT MOC375 (%)	PLANT MOC376 (%)	PLANT MOC377 (%)	PLANT MOC378 (%)	PLANT MOC379 (%)	PLANT MOC380 (%)	PLANT MOC381 (%)	PLANT MOC382 (%)	PLANT MOC383 (%)	PLANT MOC384 (%)	PLANT MOC385 (%)	PLANT MOC386 (%)	PLANT MOC387 (%)	PLANT MOC388 (%)	PLANT MOC389 (%)	PLANT MOC390 (%)	PLANT MOC391 (%)	PLANT MOC392 (%)	PLANT MOC393 (%)	PLANT MOC394 (%)	PLANT MOC395 (%)	PLANT MOC396 (%)	PLANT MOC397 (%)	PLANT MOC398 (%)	PLANT MOC399 (%)	PLANT MOC400 (%)	PLANT MOC401 (%)	PLANT MOC402 (%)	PLANT MOC403 (%)	PLANT MOC404 (%)	PLANT MOC405 (%)	PLANT MOC406 (%)	PLANT MOC407 (%)	PLANT MOC408 (%)	PLANT MOC409 (%)	PLANT MOC410 (%)	PLANT MOC411 (%)	PLANT MOC412 (%)	PLANT MOC413 (%)	PLANT MOC414 (%)	PLANT MOC415 (%)	PLANT MOC416 (%)	PLANT MOC417 (%)	PLANT MOC418 (%)	PLANT MOC419 (%)	PLANT MOC420 (%)	PLANT MOC421 (%)	PLANT MOC422 (%)	PLANT MOC423 (%)	PLANT MOC424 (%)	PLANT MOC425 (%)	PLANT MOC426 (%)	PLANT MOC427 (%)	PLANT MOC428 (%)	PLANT MOC429 (%)	PLANT MOC430 (%)	PLANT MOC431 (%)	PLANT MOC432 (%)	PLANT MOC433 (%)	PLANT MOC434 (%)	PLANT MOC435 (%)	PLANT MOC436 (%)	PLANT MOC437 (%)	PLANT MOC438 (%)	PLANT MOC439 (%)	PLANT MOC440 (%)	PLANT MOC441 (%)	PLANT MOC442 (%)	PLANT MOC443 (%)	PLANT MOC444 (%)	PLANT MOC445 (%)	PLANT MOC446 (%)	PLANT MOC447 (%)	PLANT MOC448 (%)	PLANT MOC449 (%)	PLANT MOC450 (%)	PLANT MOC451 (%)	PLANT MOC452 (%)	PLANT MOC453 (%)	PLANT MOC454 (%)	PLANT MOC455 (%)	PLANT MOC456 (%)	PLANT MOC457 (%)	PLANT MOC458 (%)	PLANT MOC459 (%)	PLANT MOC460 (%)	PLANT MOC461 (%)	PLANT MOC462 (%)	PLANT MOC463 (%)	PLANT MOC464 (%)	PLANT MOC465 (%)	PLANT MOC466 (%)	PLANT MOC467 (%)	PLANT MOC468 (%)	PLANT MOC469 (%)	PLANT MOC470 (%)	PLANT MOC471 (%)	PLANT MOC472 (%)	PLANT MOC473 (%)	PLANT MOC474 (%)	PLANT MOC475 (%)	PLANT MOC476 (%)	PLANT MOC477 (%)	PLANT MOC478 (%)	PLANT MOC479 (%)	PLANT MOC480 (%)	PLANT MOC481 (%)	PLANT MOC482 (%)	PLANT MOC483 (%)	PLANT MOC484 (%)	PLANT MOC485 (%)	PLANT MOC486 (%)	PLANT MOC487 (%)	PLANT MOC488 (%)	PLANT MOC489 (%)	PLANT MOC490 (%)	PLANT MOC491 (%)	PLANT MOC492 (%)	PLANT MOC493 (%)	PLANT MOC494 (%)	PLANT MOC495 (%)	PLANT MOC496 (%)	PLANT MOC497 (%)	PLANT MOC498 (%)	PLANT MOC499 (%)	PLANT MOC500 (%)	PLANT MOC501 (%)	PLANT MOC502 (%)	PLANT MOC503 (%)	PLANT MOC504 (%)	PLANT MOC505 (%)	PLANT MOC506 (%)	PLANT MOC507 (%)	PLANT MOC508 (%)	PLANT MOC509 (%)	PLANT MOC510 (%)	PLANT MOC511 (%)	PLANT MOC512 (%)	PLANT MOC513 (%)	PLANT MOC514 (%)	PLANT MOC515 (%)	PLANT MOC516 (%)	PLANT MOC517 (%)	PLANT MOC518 (%)	PLANT MOC519 (%)	PLANT MOC520 (%)	PLANT MOC521 (%)	PLANT MOC522 (%)	PLANT MOC523 (%)	PLANT MOC524 (%)	PLANT MOC525 (%)	PLANT MOC526 (%)	PLANT MOC527 (%)	PLANT MOC528 (%)	PLANT MOC529 (%)	PLANT MOC530 (%)	PLANT MOC531 (%)	PLANT MOC532 (%)	PLANT MOC533 (%)	PLANT MOC534 (%)	PLANT MOC535 (%)	PLANT MOC536 (%)	PLANT MOC537 (%)	PLANT MOC538 (%)	PLANT MOC539 (%)	PLANT MOC540 (%)	PLANT MOC541 (%)	PLANT MOC542 (%)	PLANT MOC543 (%)	PLANT MOC544 (%)	PLANT MOC545 (%)	PLANT MOC546 (%)	PLANT MOC547 (%)	PLANT MOC548 (%)	PLANT MOC549 (%)	PLANT MOC550 (%)	PLANT MOC551 (%)	PLANT MOC552 (%)	PLANT MOC553 (%)	PLANT MOC554 (%)	PLANT MOC555 (%)	PLANT MOC556 (%)	PLANT MOC557 (%)	PLANT MOC558 (%)	PLANT MOC559 (%)	PLANT MOC560 (%)	PLANT MOC561 (%)	PLANT MOC562 (%)	PLANT MOC563 (%)	PLANT MOC564 (%)	PLANT MOC565 (%)	PLANT MOC566 (%)	PLANT MOC567 (%)	PLANT MOC568 (%)	PLANT MOC569 (%)	PLANT MOC570 (%)	PLANT MOC571 (%)	PLANT MOC572 (%)	PLANT MOC573 (%)	PLANT MOC574 (%)	PLANT MOC575 (%)	PLANT MOC576 (%)	PLANT MOC577 (%)	PLANT MOC578 (%)	PLANT MOC579 (%)	PLANT MOC580 (%)	PLANT MOC581 (%)	PLANT MOC582 (%)	PLANT MOC583 (%)	PLANT MOC584 (%)	PLANT MOC585 (%)	PLANT MOC586 (%)	PLANT MOC587 (%)	PLANT MOC588 (%)	PLANT MOC589 (%)	PLANT MOC590 (%)	PLANT MOC591 (%)	PLANT MOC592 (%)	PLANT MOC593 (%)	PLANT MOC594 (%)	PLANT MOC595 (%)	PLANT MOC596 (%)	PLANT MOC597 (%)	PLANT MOC598 (%)	PLANT MOC599 (%)	PLANT MOC600 (%)	PLANT MOC601 (%)	PLANT MOC602 (%)	PLANT MOC603 (%)	PLANT MOC604 (%)	PLANT MOC605 (%)	PLANT MOC606 (%)	PLANT MOC607 (%)	PLANT MOC608 (%)	PLANT MOC609 (%)	PLANT MOC610 (%)	PLANT MOC611 (%)	PLANT MOC612 (%)	PLANT MOC613 (%)	PLANT MOC614 (%)	PLANT MOC615 (%)	PLANT MOC616 (%)	PLANT MOC617 (%)	PLANT MOC618 (%)	PLANT MOC619 (%)	PLANT MOC620 (%)	PLANT MOC621 (%)	PLANT MOC622 (%)	PLANT MOC623 (%)	PLANT MOC624 (%)	PLANT MOC625 (%)	PLANT MOC626 (%)	PLANT MOC627 (%)	PLANT MOC628 (%)	PLANT MOC629 (%)	PLANT MOC630 (%)	PLANT MOC631 (%)	PLANT MOC632 (%)	PLANT MOC633 (%)	PLANT MOC634 (%)	PLANT MOC635 (%)	PLANT MOC636 (%)	PLANT MOC637 (%)	PLANT MOC638 (%)	PLANT MOC639 (%)	PLANT MOC640 (%)	PLANT MOC641 (%)	PLANT MOC642 (%)	PLANT MOC643 (%)	PLANT MOC644 (%)	PLANT MOC645 (%)	PLANT MOC646 (%)	PLANT MOC647 (%)	PLANT MOC648 (%)	PLANT MOC649 (%)	PLANT MOC650 (%)	PLANT MOC651 (%)	PLANT MOC652 (%)	PLANT MOC653 (%)	PLANT MOC654 (%)	PLANT MOC655 (%)	PLANT MOC656 (%)	PLANT MOC657 (%)	PLANT MOC658 (%)	PLANT MOC659 (%)	PLANT MOC660 (%)	PLANT MOC661 (%)	PLANT MOC662 (%)	PLANT MOC663 (%)	PLANT MOC664 (%)	PLANT MOC665 (%)	PLANT MOC666 (%)	PLANT MOC667 (%)	PLANT MOC668 (%)	PLANT MOC669 (%)	PLANT MOC670 (%)	PLANT MOC671 (%)	PLANT MOC672 (%)	PLANT MOC673 (%)	PLANT MOC674 (%)	PLANT MOC675 (%)	PLANT MOC676 (%)	PLANT MOC677 (%)	PLANT MOC678 (%)	PLANT MOC679 (%)	PLANT MOC680 (%)	PLANT MOC681 (%)	PLANT MOC682 (%)	PLANT MOC683 (%)	PLANT MOC684 (%)	PLANT MOC685 (%)	PLANT MOC686 (%)	PLANT MOC687 (%)	PLANT MOC688 (%)	PLANT MOC689 (%)	PLANT MOC690 (%)	PLANT MOC691 (%)	PLANT MOC692 (%)	PLANT MOC693 (%)	PLANT MOC694 (%)	PLANT MOC695 (%)	PLANT MOC696 (%)	PLANT MOC697 (%)	PLANT MOC698 (%)	PLANT MOC699 (%)	PLANT MOC700 (%)	PLANT MOC701 (%)	PLANT MOC702 (%)	PLANT MOC703 (%)	PLANT MOC704 (%)	PLANT MOC705 (%)	PLANT MOC706 (%)	PLANT MOC707 (%)	PLANT MOC708 (%)	PLANT MOC709 (%)	PLANT MOC710 (%)	PLANT MOC711 (%)	PLANT MOC712 (%)	PLANT MOC713 (%)	PLANT MOC714 (%)	PLANT MOC715 (%)	PLANT MOC716 (%)	PLANT MOC717 (%)	PLANT MOC718 (%)	PLANT MOC719 (%)	PLANT MOC720 (%)	PLANT MOC721 (%)	PLANT MOC722 (%)	PLANT MOC723 (%)	PLANT MOC724 (%)	PLANT MOC725 (%)	PLANT MOC726 (%)	PLANT MOC727 (%)	PLANT MOC728 (%)	PLANT MOC729 (%)	PLANT MOC730 (%)	PLANT MOC731 (%)	PLANT MOC732 (%)	PLANT MOC733 (%)	PLANT MOC734 (%)	PLANT MOC735 (%)	PLANT MOC736 (%)	PLANT MOC737 (%)	PLANT MOC738 (%)	PLANT MOC739 (%)	PLANT MOC740 (%)	PLANT MOC741 (%)	PLANT MOC742 (%)	PLANT MOC743 (%)	PLANT MOC744 (%)	PLANT MOC745 (%)	PLANT MOC746 (%)	PLANT MOC747 (%)	PLANT MOC748 (%)	PLANT MOC749 (%)	PLANT MOC750 (%)	PLANT MOC751 (%)	PLANT MOC752 (%)	PLANT MOC753 (%)	PLANT MOC754 (%)	PLANT MOC755 (%)	PLANT MOC756 (%)	PLANT MOC757 (%)
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To: C. W. Klaus, Director
Environmental Protection Agency
Bureau of Water Pollution Control
625 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

Comments	PLANT PROCESSES		Boiler Plant		Boiler Plant (GPM)	REFILLABLE SOLID					ALKALINITY			LABOR		REMARKS			
	DATE	TIME	Plant	Boiler Plant		1	2	3	4	5	6	7	8	9	10				
May 1971	1	5	1400.7	17,500	2100	16,200	15	—	—	—	—	—	—	—	—	—	—	—	—
	2	5					—	—	—	—	—	—	—	—	—	—	—	—	—
	3	4	150.14	28,000	2100	26,700	5	—	—	—	—	—	—	—	—	—	—	—	—
	4	4	14,487	19,300	2400	17,700	15	—	—	—	—	—	—	—	—	—	—	—	—
	5	4	14,587	18,800	2100	17,500	2	—	—	—	—	—	—	—	—	—	—	—	—
	6	4	13,141	19,800	2100	18,500	18	—	—	—	—	—	—	—	—	—	—	—	—
	7	4	13,654	18,800	1800	17,800	2	—	—	—	—	—	—	—	—	—	—	—	—
	8	5	11,709	19,000	2100	17,700	15	—	—	—	—	—	—	—	—	—	—	—	—
	9	5					—	—	—	—	—	—	—	—	—	—	—	—	—
	10	4	16,785	24,100	2100	23,600	—	—	—	—	—	—	—	—	—	—	—	—	—
Rain	11	4	15,958	19,300	2400	17,700	16	—	—	—	—	—	—	—	—	—	—	—	—
	12	4	15,457	19,300	2100	18,000	15	—	—	—	—	—	—	—	—	—	—	—	—
	13	4	11,087	21,000	2100	19,700	1	—	—	—	—	—	—	—	—	—	—	—	—
	14	4	14,456	21,000	1800	20,000	8	—	—	—	—	—	—	—	—	—	—	—	—
	15	5	10,774	19,300	2100	18,000	18	—	—	—	—	—	—	—	—	—	—	—	—
	16	5					—	—	—	—	—	—	—	—	—	—	—	—	—
	17	4	15,533	24,700	2100	23,400	—	—	—	—	—	—	—	—	—	—	—	—	—
	18	4	16,573	21,000	2400	19,400	1	34	—	—	—	—	—	—	—	—	—	—	—
	19	4	11,231	21,900	2100	20,600	15	40	—	—	—	—	—	—	—	—	—	—	—
	20	4	9,853	19,700	2100	18,400	2	40	—	—	—	—	—	—	—	—	—	—	—
	21	4	14,691	19,500	1800	18,500	3	40	—	—	—	—	—	—	—	—	—	—	—
	22	5	10,402	12,900	2100	11,600	25	75	—	—	—	—	—	—	—	—	—	—	—
	23	5					—	—	—	—	—	—	—	—	—	—	—	—	—
	24	4	15,699	24,500	2100	23,200	1	80	—	—	—	—	—	—	—	—	—	—	—
	25	4	18,377	21,000	2100	19,400	15	85	—	—	—	—	—	—	—	—	—	—	—
	26	4	18,718	20,700	2100	19,400	1	90	—	—	—	—	—	—	—	—	—	—	—
	27	4	8,854	20,400	2100	19,100	13	95	—	—	—	—	—	—	—	—	—	—	—
	28	5	13,084	17,500	1800	16,300	1	80	—	—	—	—	—	—	—	—	—	—	—
	29	5	12,051	20,400	2100	19,100	1	70	—	—	—	—	—	—	—	—	—	—	—
	30	5					—	—	—	—	—	—	—	—	—	—	—	—	—
Total	31	4	9,789	43,800	2100	42,300	—	75	—	—	—	—	—	—	—	—	—	—	—
	32	4	346,003	582,000	54,000	528,200	348	804	—	—	—	—	—	—	—	—	—	—	—
	33	4	13,308	18,129	1761	17,039	15	67	—	—	—	—	—	—	—	—	—	—	—
	34	4	18,718	43,800	2400	27,200	3	95	—	—	—	—	—	—	—	—	—	—	—
	35	4	8,141	12,900	1800	11,600	5	34	—	—	—	—	—	—	—	—	—	—	—
4 (Well - Boiler + 800 GPM TAP)																	4 (Solidity Flow 1400, 14 Monday's)		

**To: C. W. Klausen, Director
Environmental Protection Agency
Bureau of Water Pollution Control
525 W. Jefferson St.
Springfield, Illinois**

MONTHLY OPERATION REPORT

MONTHLY OPERATION REPORT																		
Comments	PLANT PROCESS			Plant	Boiler Fuel (Gals.)	BETTERABLE SOLDS			ALKALINITY			WATER						
	DATE	TIME	IN.			1	2	3	1	2	3							
JUNE 1971	1	T	16,013	16,800	2,100	15,100	2	75	6.5	7.8	7.4	260	520	376	0.63	0.63		
RAIN	2	W	12,043	16,800	2,100	15,500	2	80	6.4	7.4	7.4	270	520	356	0.64	0.64		
	3	T	7,129	16,700	2,100	15,400	2.5	85	6.3	7.4	7.5	300	516	346	0.64	0.64		
Worst Blower out of order this A.M.	4	F	6,888	21,900	1800	20,900	1	80	6.2	7.4	7.4	260	526	420	0.64	0.64		
Worst Blower out of order this A.M.	5	S	11,658	24,800	2,100	23,500	1	95	6.3	7.4	7.5	300	540	430	0.64	0.64		
RAIN	7	M	16,782	41,400	2,100	43,100	—	130	6.4	7.5	7.6	200	540	420	0.70	0.70		
	8	T	8,792	20,600	2,100	19,500	2	140	6.4	7.4	7.6	180	530	370	0.70	0.70		
	9	W	13,571	19,700	2,100	19,400	1.5	130	6.3	7.5	7.5	186	536	376	0.70	0.70		
	10	T	8,691	21,900	2,100	20,600	2	140	6.3	7.5	7.5	260	540	370	0.71	0.71		
	11	F	14,755	16,700	1800	15,700	1.5	130	6.3	7.5	7.5	240	540	370	0.71	0.71		
	12	S	10,083	16,700	2,100	15,400	1	130	6.3	7.4	7.6	200	536	370	0.70	0.71		
WE ARE RUNNING THROUGH WITH HIGH TEMPERATURES ON OUR REFRIGERATORS. THEY ARE BECOMING OFF OF THE LINE AIR TO LOWER REFRIGERATION.	13	S	14,645	24,500	2,100	22,600	—	110	6.4	7.4	7.5	180	530	376	0.71	0.71		
	14	M	15,520	22,200	2,100	20,600	1.5	130	6.3	7.4	7.4	206	536	374	0.71	0.71		
	15	T	12,301	23,200	2,100	21,900	2	135	6.3	7.4	7.5	240	526	372	0.71	0.71		
	16	T	8,584	18,800	2,100	17,500	1.5	130	6.3	7.5	7.6	240	530	346	0.71	0.71		
	17	F	14,184	23,700	1800	22,100	2	135	6.3	7.4	7.6	190	526	370	0.72	0.72		
	18	S	12,404	20,600	2,100	19,300	1.2	130	6.2	7.5	7.5	210	526	370	0.71	0.71		
RAIN	21	M	14,916	34,200	2,100	32,700	—	120	6.3	7.6	7.6	140	500	380	0.72	0.72		
	22	T	13,270	20,100	2,100	18,300	1.5	120	6.3	7.5	7.6	280	530	412	0.74	0.74		
	23	W	13,486	19,700	2,100	18,400	1.5	120	6.2	7.5	7.6	240	526	410	0.75	0.75		
	24	T	8,055	18,800	2,100	17,500	1	120	6.3	7.6	7.5	260	530	416	0.74	0.74		
	25	F	15,152	21,500	1800	20,500	1.5	120	6.3	7.5	7.6	200	530	416	0.74	0.74		
	26	S	11,750	16,700	2,100	15,400	1.5	120	6.2	7.5	7.5	212	530	420	0.76	0.76		
	27	S	15,563	27,800	2,100	26,500	—	120	6.3	7.6	7.6	180	426	416	0.75	0.75		
	28	M	11,883	29,300	2,100	27,100	1.5	120	6.3	7.0	7.4	200	4400	3800	0.80	0.80		
	29	W	11,852	27,200	2,100	25,900	2	120	6.3	7.3	7.5	216	440	310	0.81	0.81		
	30	T	330	640	58720	5400	553100	31.4	306	—	104	194	155	5870	13,510	10,102	0.82	0.82
	31	F	12,725	19,573	1830	18,437	1.5	118	6.3	7.4	7.5	227	520	387	0.72	0.72		
	1	S	16,828	44,400	2,400	43,100	2.3	140	6.5	7.6	7.6	300	540	430	0.81	0.81		
	2	S	7,129	16,700	1800	15,100	1.2	75	6.2	7.0	7.4	140	426	310	0.66	0.66		
* (WELL - Boiler + BOC GPD Inlet)																		
** (SUPPORT PUMP Inlet in MONTHS)																		

To: C. W. Eklund, Director
Environmental Protection Agency
Bureau of Water Pollution Control
525 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

Comments	DATE	PLANT PROCESS		Boiler Flow (GAL.)	Boiler Flow (GAL.)	SETTABLE SOLIDS			ALKALINITY			LABOR		WEATHER
		DATE	CHARGE PPTS. PER HOUR	PER HOUR	PER HOUR	1	2	3	1	2	3	1	2	
JULY 1971	1		6,175	27,200	2,500	1.2	100	—	6.3	7.1	7.5	400	460	
	2	F	15,180	21,500	1900	1	120	—	6.2	7.1	7.5	536	440	0.80.580
	3	S	5,075	27,200	2,300	1.3	120	—	6.3	7.1	7.5	520	444	0.75.15.74
	4	S												0.74.15.75
RAIN	5	M	14,863	37,400	1900	—	120	—	6.2	7.5	7.5	516	440	0.77.18.75
	6	T	13,628	23,200	2,100	2	100	—	6.3	7.1	7.6	500	420	0.74.28.74
	7	W	15,146	23,400	2,100	2	50	—	6.3	7.1	7.6	506	416	0.74.30.75
	8	T	7,517	25,000	2,200	2	50	—	6.2	7.1	7.5	500	400	0.75.50.75
	9	F	15,234	24,500	1700	2	50	—	6.3	7.1	7.5	500	360	0.73.20.75
	10	S	12,773	21,900	2,100	1.5	40	—	6.2	7.1	7.5	500	330	0.72.16.74
	11	S												
	12	M	14,177	42,900	2,200	—	50	—	6.3	7.3	7.4	160	400	0.72.25.72
	13	T	14,186	25,000	3,300	.8	50	—	6.3	7.5	7.1	210	410	0.72.15.72
	14	W	12,569	25,000	2,300	.6	50	—	6.3	7.2	7.2	216	416	0.76.25.73
RAIN	15	T	8,588	24,100	1800	.5	50	—	6.3	7.2	7.2	246	420	0.72.37.73
	16	F	11,553	27,200	1800	.2	50	—	6.3	7.2	7.2	240	480	0.73.35.74
	17	S	10,173	22,800	1900	0	50	—	6.2	7.2	7.2	200	410	0.72.35.73
	18	S												
	19	M	14,777	30,400	2,600	—	—	—	6.3	7.2	7.3	240	520	2.5.72.27.70
	20	T	11,574	25,000	2,300	—	—	—	6.2	7.3	7.1	240	510	4.0.70.35.70
	21	W	10,333	25,800	1700	—	—	—	6.2	7.1	7.2	230	520	4.0.71.37.71
	22	T	7,706	20,600	1500	.2	—	—	6.2	7.1	7.1	240	516	4.0.71.35.72
	23	F	17,949	24,100	1900	.2	—	—	6.3	7.2	7.2	200	520	1.5.71.30.72
	24	S	8,775	24,300	2,400	1.5	—	—	6.2	7.1	7.2	280	512	2.70.35.71
RAIN	25	S												
	26	M	14,340	40,300	1500	—	—	—	6.2	7.1	7.2	160	500	1.0.67.30.70
	27	T	13,236	25,800	2800	1.2	—	—	6.2	7.2	7.2	240	500	0.49.50.70
	28	W	12,386	21,500	2,100	1	—	—	6.3	7.1	7.2	200	506	2.49.52.71
	29	T	7,590	20,600	1800	.6	—	—	6.2	7.1	7.2	260	500	2.70.60.71
	30	F	16,743	20,600	2,100	.2	—	—	6.1	7.2	7.1	260	510	0.70.55.71
	31	S	12,549	19,200	2,100	1.5	—	—	6.2	7.2	7.2	280	500	.5.49.50.70
	Total		311,435	706,800	50,000	25	1050	—	6.26	7.58	7.75	4714	1364	1.7.71.55.74
	Avg.		11,535	22,800	1610	1.0	70	—	6.2	7.3	7.3	229	504	.7.72.31.73
	Max.		17,949	42,500	3300	2	120	—	6.3	7.5	7.4	280	536	4.80.60.70
	Min.		5,075	19,700	1500	0	40	—	6.1	7.1	7.1	160	480	0.49.57.70
* (SUNDAY) Flow Incl. in Monday's														
* (WELL-Boiler + BOD CAPD INFLU.)														

To: C. W. Klaman, Director
Environmental Protection Agency
Bureau of Water Pollution Control
1215 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

[illegible]

To: C. W. Klassen, Director
Environmental Protection Agency
Bureau of Water Pollution Control
525 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

[illegible]

To: C. W. Klassen, Director
Environmental Protection Agency
Bureau of Water Pollution Control
525 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

[illegible]

**To: C. W. Klassen, Director
Environmental Protection Agency
Bureau of Water Pollution Control
525 W. Jefferson St.
Springfield, Illinois**

MONTHLY OPERATION REPORT

Springfield, Illinois

Comments

NOVEMBER 1971

RAIN

DATE	DAY	PLANT PROCESS		CHINA PITCH INCH	Plant TIME (HRS)	FLOOR INCH	LACONIA INCH	BETTERABLE			ALCALINITY			LACONIA	RELATIVE P
		TIME	REACT.					1	2	3	1	2	3		
1	M			17.747	17.747	24,750	2300	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
2	T			21.450	21.450	2300	17,300	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
3	W			17.451	17.451	2000	15,400	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
4	T			15.121	15.121	2300	18,600	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
5	F			11.241	11.241	1300	11,300	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
6	S			14.623	14.623	1400	15,400	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
7	S														
8	M			17.671	17.671	26000	24,000	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
9	T			13.105	13.105	18,900	3300	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
10	W			14.815	14.815	14,600	2000	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
11	T			12.225	12.225	13,400	2000	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
12	F			14.103	14.103	8,600	1700	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
13	S			17.154	17.154	9,000	18,000	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
14	S														
15	M			20.313	20.313	24,000	24,000	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
16	T			20.656	20.656	14,100	2500	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
17	W			23.264	23.264	18,150	2500	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
18	T			14.963	14.963	2100	15,313	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
19	F			21.943	21.943	16,215	2000	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
20	S			16.073	16.073	12,871	1300	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
21	S					19,000	2600	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
22	M			19.657	19.657	16,788	2500	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
23	T			16.470	16.470	18,000	15,412	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
24	W			10.174	10.174	16,000	1100	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
25	T					12,100	2300	2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
26	F			16.394	16.394			2.5	2.5	2.5	2.5	2.5	2.5	2.5	2.5
27	S			19.315	19.315	22,815	2100	21,575							
28	S			19.925	19.925	26000	17,915								
29	M			14.430	14.430	17,500	18,000	16,500							
30	T			19.042	19.042	18,000	16,000	17,000							
31	W			416.845	416.845	413,000	56,800	481,600							
TOTAL				15,454	15,454	15,710	18,913	14,587							
Aug.				23,264	23,264	38,000	33,000	24,000							
Nov.				10,174	10,174	8,600	11,000	7,700							
Dec.															
				1972											
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To: C. W. Klamm, Director
Environmental Protection Agency
Bureau of Water Pollution Control
1225 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

Comments	DATE	DAY	PLANT PROCESS		Plant Flow (MGD)	Boiler Flow (MGD)	SETTLABLE SOLIDS		pH			ALKALINITY			JANUARY		FEBRUARY		WEATHER		
			Run	Perf.			1	2	3	1	2	3	1	2	3	1	2	1	2	1	2
DECEMBER 1971					11,495	2340	15,935			6.2	7.2	7.2	240	490	430	5	10	5	10	5	10
CHARGE CALIBRATION OF EFF.	1	W			23,345	2400	21,505			6.2	7.2	7.2	230	520	400	8	35	8	35	8	35
"	2	F			17,245	2340	15,745			6.2	7.2	7.2	230	520	462	5	39	11	36	5	39
"	3	S			21,750	2800	19,750			6.2	7.2	7.2	240	520	408	4	40	11	38	4	40
INFILTRATION TEST ON SENSER LINE	4	S			9,875	1700	8,925			6.2	7.2	7.2	180	520	480	4	40	8	38	4	40
CHARGE CALIBRATION OF EFF.	5	M			17,540	2700	15,440			6.2	7.2	7.2	240	520	480	3	41	11	41	3	41
RAIN	6	W			21,210	2700	19,310			6.2	7.2	7.1	240	522	480	12	41	11	41	12	41
"	7	T			15,545	2000	14,345			6.2	7.2	7.2	240	520	476	4	44	13	43	4	44
"	8	F			16,765	2300	15,265			6.2	7.2	7.2	240	520	476	15	45	13	43	15	45
CHARGE CALIBRATION OF EFF.	9	F			15,980	2300	14,480			6.2	7.2	7.2	240	520	476	5	44	13	43	5	44
INFILTRATION TEST ON SENSER LINE	10	S			11,450	1600	12,650			6.2	7.1	7.2	230	516	480	5	44	13	43	5	44
"	11	S			9,875	1000	9,675			6.2	7.2	7.2	194	520	480	6	45	13	45	6	45
"	12	M			17,805	2000	15,805			6.2	7.2	7.2	190	520	460	5	40	14	41	5	40
"	13	W			17,540	2700	15,640			6.2	7.2	7.2	230	516	460	5	45	14	41	5	45
"	14	T			17,855	2500	16,155			6.2	7.2	7.2	220	520	460	3	41	14	41	3	41
"	15	F			15,835	2400	14,235			6.2	7.2	7.2	220	500	440	3	41	14	41	3	41
"	16	S			13,270	2300	11,770			6.2	7.2	7.2	220	500	440	3	41	14	41	3	41
INFILTRATION TEST ON SENSER LINE	17	S			13,075	2000	11,875			6.2	7.2	7.2	220	500	444	4	43	15	41	4	43
"	18	M			11,125	1600	10,325			6.2	7.2	7.2	200	500	456	15	37	10	35	15	37
"	19	T			20,640	2500	18,940			6.2	7.2	7.2	240	500	460	30	35	11	35	30	35
"	20	W			18,570	2400	16,970			6.2	7.2	7.2	240	480	490	35	35	11	35	35	35
"	21	T			15,975	2200	14,575			6.2	7.2	7.2	230	484	480	35	37	11	35	35	37
"	22	F			19,105	2500	17,605			6.2	7.2	7.2	240	490	484	40	36	11	35	40	36
"	23	S			25,195	2500	23,495			6.2	7.2	7.2	200	484	470	40	36	12	36	40	36
"	24	M			13,215	1500	12,515			6.2	7.2	7.2	230	484	476	35	37	11	35	35	37
"	25	T			18,800	3000	16,800			6.2	7.2	7.2	240	490	480	20	37	10	36	20	37
"	26	W			20,955	3000	18,955			6.2	7.2	7.2	240	490	484	5	36	9	36	5	36
"	27	T			20,435	3000	18,435			6.2	7.2	7.2	240	492	484	2	36	9	36	2	36
"	28	F			20,120	2400	18,520			6.2	7.2	7.2	230	490	484	3	36	9	36	3	36
"	29	S			50,505	4000	46,505			6.2	7.2	7.2	240	13,020	12,714	7	10	3	10	7	10
"	30	M			16,340	2223	14,811			6.2	7.2	7.2	227	505	471	3	36	9	36	3	36
"	31	T			23,185	3100	23,185			6.2	7.2	7.2	240	522	490	8	45	15	45	8	45
"					9,155	1000	8,925			6.2	7.1	7.1	180	480	440	2	33	9	34	2	33
P(Well-Boiler + 1800 GPD TREAT)																					

To: C. W. Klausen, Director
Environmental Protection Agency
Bureau of Water Pollution Control
525 W. Jefferson St.
Springfield, Illinois

MONTHLY OPERATION REPORT

[illegible]

APPENDIX D
Cost Estimates

Annuity whose present value is 1

$$A_{\overline{n}|i}^{-1} = 1 / (1 - v^n) = S_{\overline{n}|i}^{-1} + i$$

Mechanical Equipment - \$15,592.82

$$n = 5 \text{ yr } i = 0.075 \text{ (7 1/2\%)} \quad 0.24716472 \times \$15,592.82 = 3853.99$$

Construction - \$33,900.56

$$n = 30 \text{ yr } i = 0.075 \text{ (7 1/2\%)} \quad 0.08467124 \times 33,900.56 = 2870.40$$

where

$A_{\overline{n}|i}$ present value of annuity for unit periodic payment
at interest rate i

i = rate of interest

n = number of conversion periods

v = present value (at compound interest) for unit principal,
 $= (1 + i)^{-1}$

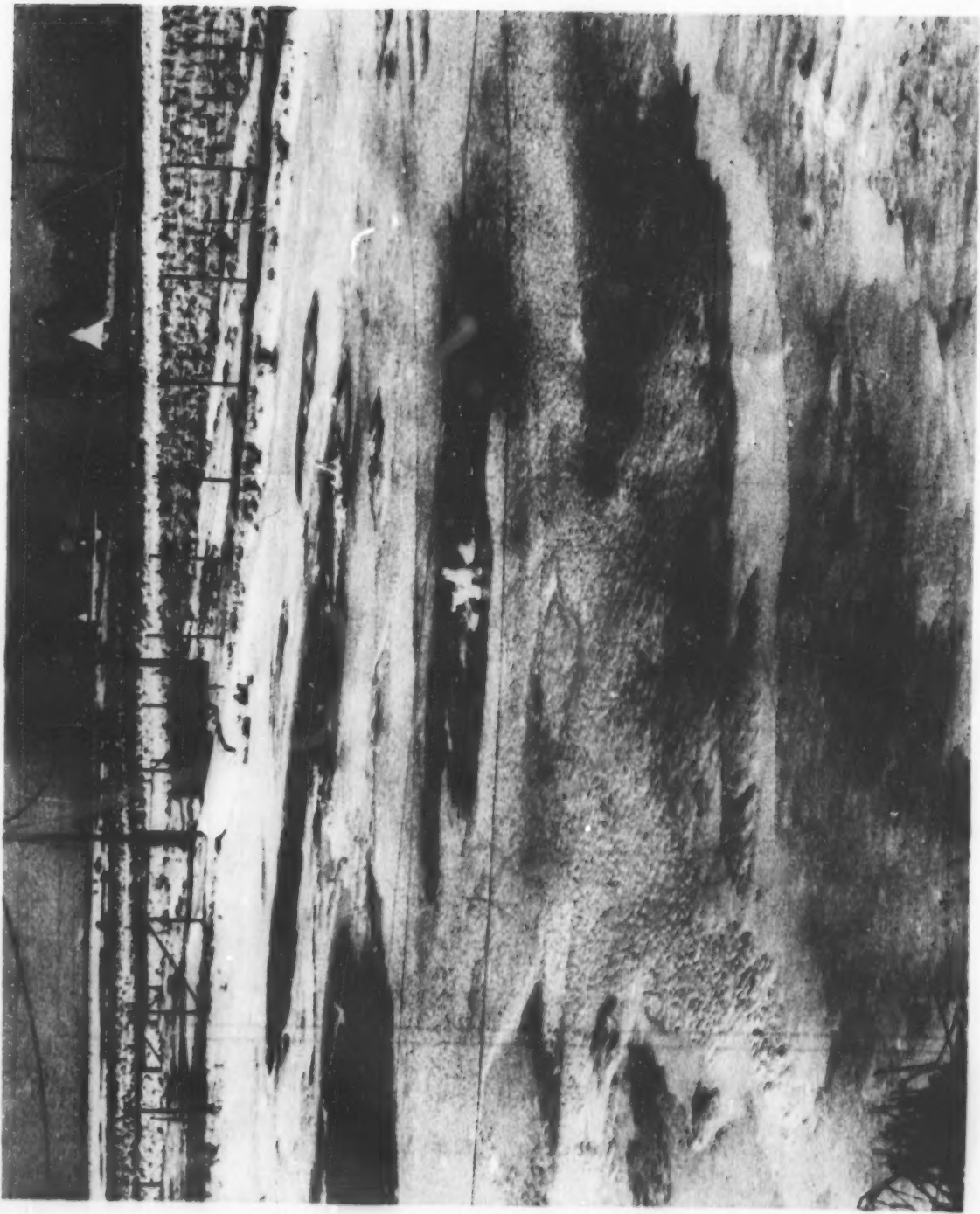
S = Amount (at compound interest) for unit principal = $(1 + i)^n$

APPENDIX E

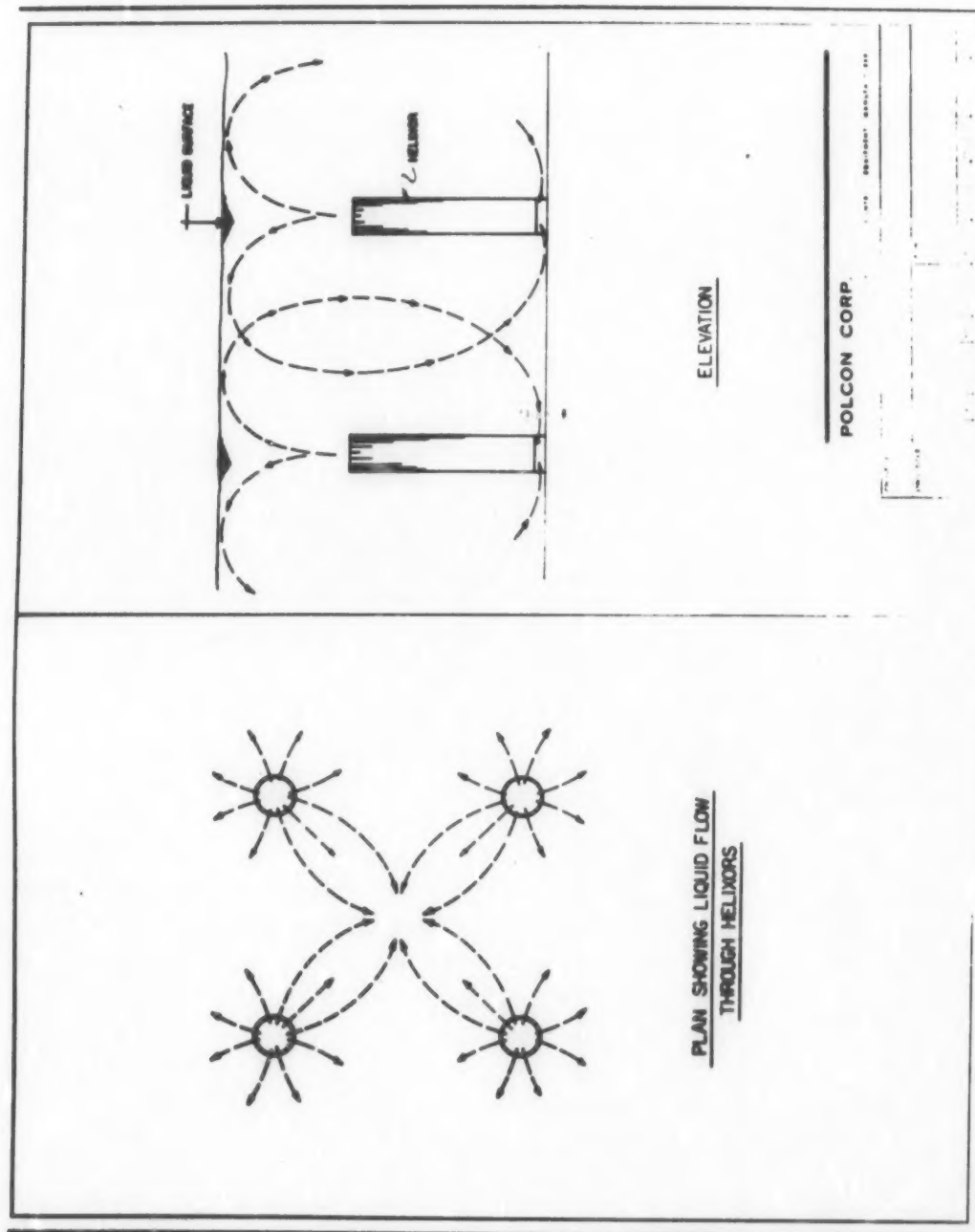
Pictures



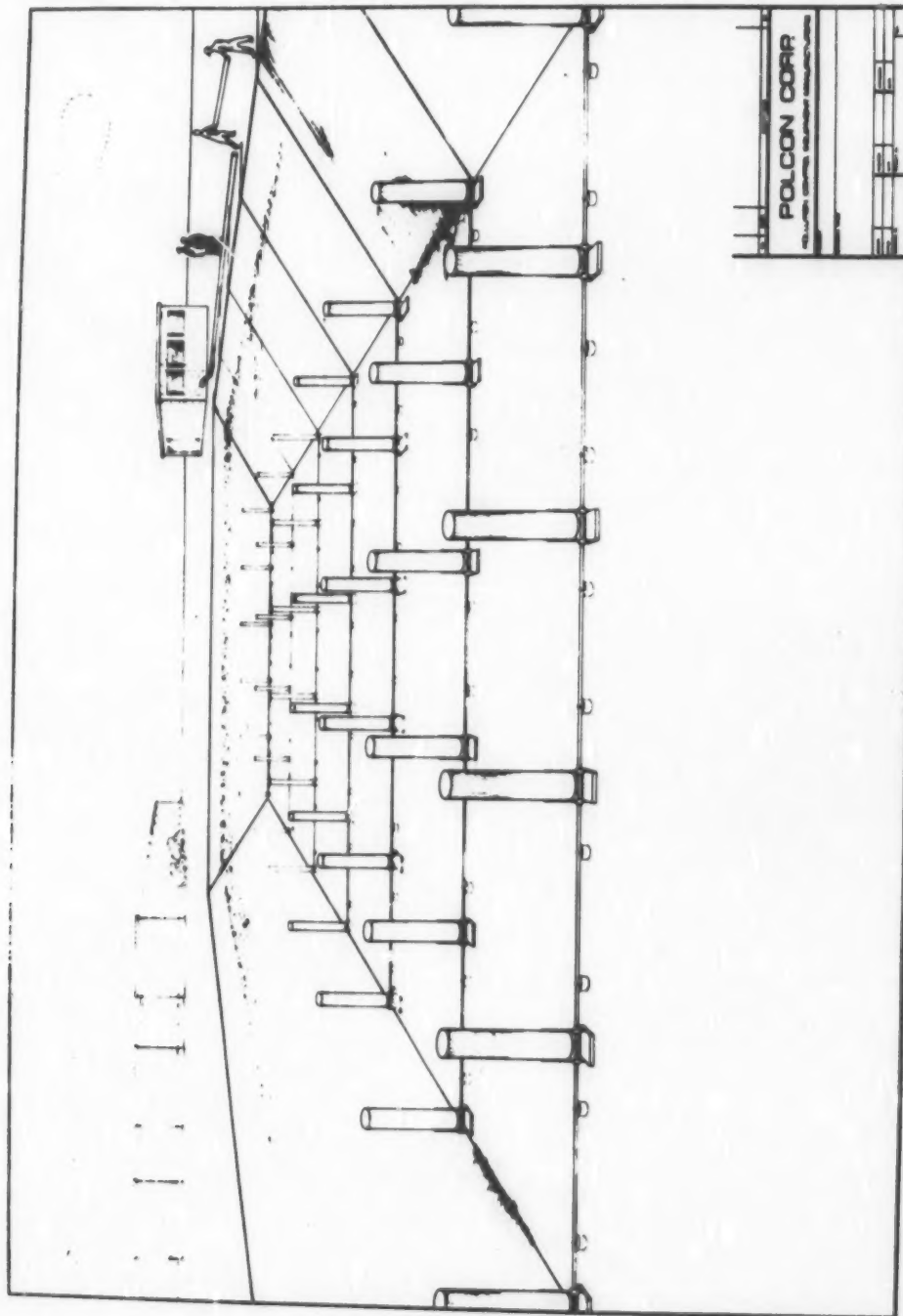
Primary Lagoon - Summer



Primary Lagoon - Winter



Polcon Helixor - Flow Diagram



Polcon Helixor - Perspective